1	Direct determination of AI and Pb in waste printed circuit boards (PCB) by
2	Laser-induced breakdown spectroscopy (LIBS): evaluation of calibration
3	strategies and economic - environmental questions
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# 30 Graphical abstract



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Evaluation of five calibration strategies for direct determination of recyclable and
 toxic metals in waste PCB by LIBS

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# 35 Highlights

- Fast and direct determination of AI and Pb in waste PCB by LIBS
- Evaluation of five calibration strategies to overcome matrix effects by LIBS
- Recent calibration strategies by LIBS were applied to determine AI and
   Pb in WEEE.
- Pb concentration are 5 to 12 times higher than established by Directive
   2011/65/EU
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49 Abstract

Matrix-matching calibration (MMC), two-point calibration transfer (TP CT), one-50 point and multi-line calibration (OP MLC), single-sample calibration (SSC) and 51 calibration free (CF) were evaluated in order to overcome matrix effects in laser-52 induced breakdown spectroscopy (LIBS). These calibration strategies were 53 evaluated for direct determination of AI and Pb in waste printed circuit boards 54 (PCB) using direct solids analysis by LIBS. Each strategy has limitations and 55 advantages of its implementation, for the correction of matrix effects, so that it 56 allows elementary determination with adequate accuracy. The MMC and CF 57 58 proved to be excellent calibration strategies for the determination of strategic (AI) and toxic (Pb) elements by LIBS, with good recoveries (ranging from 80 to 59 120%) and low relative standard deviation (RSD%) values. A detailed 60 61 discussion of the advantages and limitations of each of these five calibration strategies evaluated for LIBS is presented in this study. Lead concentrations in 62 waste PCB samples are 5 to 12 times higher than established by Directive 63 2011/65/EU, and the samples analyzed contain between 3 and 55 g kg<sup>-1</sup> Al, 64 being an interesting economic and recycling source for this metal. 65

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Keywords: WEEE; Toxic element; Environment pollution; Matrix-matching
 calibration; Two-point calibration transfer; Calibration free; Single-sample
 calibration; One-point multi-line calibration

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#### 75 **1. Introduction**

Contemporary society utilizes several types of high-tech electrical and electronic devices and instruments. The study of waste electrical and electronic equipment (WEEE) generation is therefore of great interest due to the environmental, economic, recycling and reuse questions posed by this type of waste (Costa et al. 2018a, Tansel, 2017). According to a study published in 2017, the projection of WEEE production is expressive (Andrade et al. 2019a, 2019b, 2019c, 2019d), being 52.2 million tons in 2021 (Baldé et al. 2017).

WEEE contains a diverse, complex and valuable composition, 83 consisting mainly of polymers, metals (base, toxic, noble and technological 84 elements) and ceramics. Printed circuit boards (PCB), which are part of the 85 electronic devices present in the WEEE, contain large amounts of valuable and 86 87 dangerous metals such as, for example, AI and Pb. Depending on the electronic device, the metal content on PCBs can range from 2 to 19% Al and 1 to 3% Pb, 88 among others (Andrade et al. 2019b, Arshadi et al. 2018, Carvalho et al. 2015, 89 Perkins et al. 2014, Yamane et al. 2011, Yang et al. 2019). Aluminum and Pb 90 can be recycled, adding value to the WEEE and generating a source of income. 91 92 Lead is a toxic metal which in high concentrations can pollute the environment. According to Directive 2011/65/EU, the maximum concentration allowed is 0.1% 93 by weight in homogeneous materials (RoHS, 2011). 94

The preparation of waste PCB samples for elemental analysis is a challenging task due to the complexity of its composition and its refractory character, being rich in oxides of Mg, Si and Ti, flame retardants, metals and polymers. Thus, the need to use vigorous conditions in the acid decomposition step is evident in order to obtain a homogeneous and representative solution for

subsequent quantitative analysis by conventional analytical techniques (Arshadiet al. 2018).

LIBS technique has some advantages that could be used for the direct analysis of waste PCBs, such as: minimum sample preparation, fast multielement analysis (µs), semi-destructive analysis (µg), and minimum waste generation. A limitation of this technique is related to matrix effects, which may compromise accuracy in quantitative analysis (Andrade et al. 2020, Carvalho et al. 2018a, Costa et al. 2019, Cremers and Radziemski 2006, Gondal et al. 2010, Kim et al. 2013, Lasheras et al. 2011, Miziolek et al. 2006).

109 As the sample is analyzed integrally (analyte and matrix simultaneously) by the LIBS instrument, the physicochemical properties of the 110 sample and the laser-sample and/or laser-plasma interaction may influence the 111 112 atomic/ionic/molecular emission phenomenon of the analyte (Cremers and Radziemski 2006a, Miziolek et al. 2006). As in the majority of applications the 113 114 goal is quantitative analysis, requiring calibration standards in some strategies. 115 In addition, different matrix effects can occur in the plasma formed in the samples and in the calibration standards. Consequently, the figures of merit of 116 the method can be jeopardized, and thus may make it impossible to determine 117 the analyte with satisfactory accuracy using direct solid analysis by LIBS (Hahn 118 and Omenetto 2010, Hahn and Omenetto 2012, Sattar et al. 2019). 119

Matrix effects (spectral and non-spectral) can be avoided or minimized by careful peak selection or peak fitting of the analytical line and / or selecting lines that do not exhibit spectral interference, in addition the use of high resolution spectrometers in LIBS (makes it possible to identify and overcome some spectral interferences) (Takahashi and Thornton 2017, NIST).

The non-spectral matrix effects are directly correlated to the physical 125 and chemical properties of the sample, and these effects are more difficult to 126 overcome because there are many possibilities of how the matrix may be 127 influenced by the analyte emission phenomenon (Cremers and Radziemski 128 2006a, Takahashi and Thornton 2017). The main non-spectral matrix effects 129 are related to the sample's irregular surface, inhomogeneous particle size and 130 humidity (Carvalho et al. 2018, Takahashi and Thornton 2017), the predominant 131 chemical composition of the sample (organic or inorganic forms) (Eppler et al. 132 1996), the temperature of the sample (Lednev et al. 2019), the pressure used to 133 134 compress the sample to form pellets (when necessary) (Popov et al. 2018), the presence of easily ionizable elements (EIEs) (Morais et al. 2018, Popov et al. 135 2018), the interaction of the laser with the sample, heat of vaporization, thermal 136 137 conductivity, and the absorption coefficient, which affects the transport of an ablated mass which will be vaporized and atomized into the plasma (Takahashi 138 and Thornton 2017), among others (Lasheras et al. 2013, Rezaei et al. 2018). 139 140 These matrix effects may contribute to the non-stoichiometric ablation of the sample and thus reduce the possibility of using LIBS for quantitative analysis. 141

In order to overcome these matrix effects, univariate and multivariate 142 calibration strategies are used for LIBS: matrix-matching calibration (MMC) 143 (Costa et al. 2018b, Gomes et al. 2013, Vieira et al. 2018), standard addition 144 (SA) (Yi et al. 2016, Wu et al. 2019), one-point gravimetric standard addition 145 (OP GSA) (Babos et al. 2019), internal standardization (IS) (Aquino et al. 2016, 146 Carvalho et al. 2018b, Lasheras et al. 2013, Sperança et al. 2019) calibration 147 free (CF) (Calvacante et al. 2013, Ciucci et al. 1999, Li et al. 2019, Tognoni et 148 al. 2010), one-point and multi-line calibration (OP MLC) (Hao et al. 2018), multi-149

energy calibration (MEC) (Andrade et al. 2019b, Augusto et al. 2019, Babos et
al. 2018, Carvalho et al. 2019, Castro et al. 2020, Fortunato et al. 2019), twopoint calibration transfer (TP CT) (Castro et al. 2020), and single-sample
calibration (SSC) (Yan et al. 2019).

It should be noted that many calibration possibilities are available for 154 LIBS, but the question is what is the best calibration strategy that could be 155 applied for the determination of AI and Pb in samples as complex, of 156 environmental and economic interest as waste PCBs. In order to answer this 157 question, five calibration strategies were selected and evaluated their 158 performance based on various parameters such as standard error (SE), 159 recovery and other figures of merit, considering the limitations and intrinsic 160 advantages of each calibration for overcoming matrix effects and for the 161 162 determination of these metals. Two strategies widely reported in the literature (MMC and CF) and three that have recently been proposed (TP CT, OP MLC 163 164 and SSC) for calibration were evaluated for direct waste PCB analysis and determinations of toxic (Pb) and strategic (AI) elements by LIBS. 165

166 **2. Experimental** 

## 167 2.1. Instrumentation

LIBS spectra were obtained using an experimental set-up based on a Q-switched Nd:YAG laser (BrilliantQuantel, model Ultra CFR) with a 1064 nm wavelength, a 7.7 ns pulse duration and a maximum laser pulse energy of 50 mJ. The sample was placed inside a sample chamber and the laser beam directly focused on it through a 150 mm focal length lens. The target surface was positioned approximately 77 mm below the focal lens. The light emitted by the plasma was collected by optic fibers connected to an Echelle spectrometer

(Andor Mechelle ME5000, 195 mm focal length, F/7, I/AI 5000). The 175 spectrometer is equipped with an intensified charge coupled device detector 176 (Andor iStar DH734, 1024  $\times$  1024 pixels 13.6  $\times$  13.6  $\mu$ m<sup>2</sup> by pixel, 18 mm of 177 intensifier diameter). The wavelength and spectral resolution of the 178 spectrometer were calibrated using a low pressure mercury-argon lamp by 179 measuring both the spectral positions of the lines and their spectral profiles. The 180 LIBS system requires some instrumental parameters to be optimized such as 181 laser pulse energy, delay time, signal acquisition time and lens-sample 182 distance. 183

An inductively coupled plasma optical emission spectrometer (ICP OES) (iCAP 7000, Thermo Scientific, Waltham, MA, USA) was used in the determination of AI and Pb in printed circuit board waste after acid digestion of the samples (n=3). The concentrations obtained were used as reference values for the LIBS method. The emission lines monitored during ICP OES determinations were AI 167.079 nm and Pb 216.99 nm using axial viewing mode.

A scanning electron microscope (SEM) (JEOL JSM 6360-LV) (General 191 Research Support Service of the University of Zaragoza) with voltage up to 30 192 kV and a maximum resolution of 3.0 nm was used for the morphological surface 193 visualization of the waste PCBs. In order to perform the analysis, one waste 194 PCB sample was pelletized and selected (S2). Additionally, the pellet was 195 196 analyzed by SEM in order to obtain information about the crater formed by the laser pulse and then to calculate both the irradiance and laser pulse fluence 197 values. 198

199 2.2. Reagents and samples

Standard solutions containing AI and Pb were prepared by diluting standard stock solutions containing 1000 mg L<sup>-1</sup> (Specsol, São Paulo, Brazil), and acidified with HNO<sub>3</sub> 10% v v<sup>-1</sup>, and used for ICP OES analysis.

Six PCBs from desktop computers were collected at São Carlos (São 203 Paulo State, Brazil) and then ground in a knife mill (IKA, A11) (the particle size 204 was estimated to be lower than 500 µm) to obtain homogeneous and 205 representative samples. Approximately 200 mg of PCB samples were weighed 206 207 and compressed using a manual hydraulic press (Perkin Elmer IR Accessory Hydraulic Press) with 10x10<sup>4</sup> N for 2 min, to obtain pellets (n=3) for LIBS 208 209 analysis. It is necessary to press the sample to obtain cohesive pellets that contribute to the reproducibility of the laser-sample interaction and consequently 210 211 to the precision of the measurements.

212 2.3. Sample preparation for determination of Al and Pb by ICP OES

The samples of PCBs were digested using microwave-assisted heating 213 214 for analysis and to obtain reference values of AI and Pb that were subsequently 215 used in the proposition of the calibration models and verification of the accuracy of the proposed LIBS method. Masses of approximately 100 mg of PCB were 216 accurately weighed directly in the perfluoroalkoxy alkanes (PFA) digestion 217 vessels and microwave-assisted digested using a single reaction chamber oven 218 (UltraWave<sup>™</sup>, Milestone, Sorisole, Italy). Volumes of 5 mL of concentrated 219 HNO<sub>3</sub> were used as an oxidizing agent in the decomposition. The microwave 220 heating program was applied as follows: (1) 5 min to reach 100 °C, (2) 15 min 221 to reach 180 °C, (3) 15 min to reach 240 °C and (4) 7 min held at 240 °C. 222 Subsequently, the digests were diluted to 50.0 mL with distilled-deionized water 223

and filtered on qualitative filter paper 80 g m<sup>-2</sup> (Unifil, Germany) for subsequent
ICP OES analysis.

226 2.4. Optimization of LIBS instrumental parameters

Using a full factorial design  $2^3$  with center and axial points, the 227 instrumental conditions (delay time, gate width and laser pulse energy) were 228 optimized. The variables studied were evaluated at five levels: delay time (0.01, 229 0.4, 1.2, 2.0 and 2.54 µs), gate width (0.32, 1.0, 2.0, 3.0 and 3.68 µs) and laser 230 pulse energy (20, 25, 35, 42.5 and 47.5 mJ). The variable levels were coded 231 between -1.68 (lower level) and +1.68 (higher level), with the central point 232 (coded as 0) used to calculate experimental errors. Table S1 presented at 233 supplementary material shows more details about the experimental design 234 performed. Due to experimental setup limitations the delay time values were 235 236 coded from -1.49 (0.01 µs) to 1.68 (2.54 µs) The S2 waste PCB sample containing 55  $\pm$  3 g kg<sup>-1</sup> Al and 11.6  $\pm$  0.8 g kg<sup>-1</sup> Pb was used to optimize the 237 238 instrumental conditions used in the LIBS analyses. The AI and Pb reference concentrations were obtained after microwave digestion and ICP OES 239 determinations. 240

The signal-to-background ratio (SBR) calculated for each monitored 241 emission line of AI and Pb (besides the lines of Ca, Fe, Si and Ti used for 242 calibration free) were used as responses of the factorial design. A mathematical 243 approach developed by Derringer and Suich (1980), based on desirability 244 functions applied to optimize multi-response experiments, was used in this 245 study. This strategy first converts each experimental response into an individual 246 desirability value (d<sub>i</sub>), which ranges between  $0 \le d_i \le 1$ . In this case,  $d_i = 1$ 247 corresponds to a desired response (high SBR), while  $d_i = 0$  represents a 248

response that is outside the acceptable region (the lowest SBR). The individual desirability value was combined into a single response after an arithmetic mean calculation (the overall desirability, OD). In this study, was exceptionally used the arithmetic mean because some experiments resulted in di = 0.

253 2.5. Calibration strategies

Five calibration strategies were evaluated for the determination of Al and Pb in waste PCB samples by LIBS. For all the calibration strategies, the calibration standards and samples were pelletized (n=3) using approximately 200 mg of waste PCB. The resulting pellets were analyzed using 50 pulses in different spots to obtain a single average spectrum. For each sample, six average spectra were obtained (total of 300 shots per sample).

Eight different strategies for normalization of the spectra (Castro and Pereira-Filho 2016, Sperança et al. 2018) were evaluated. These normalizations are important to minimize the signal fluctuations (area or height) and sample matrix differences during data acquisition.

264

2.5.1 Matrix-matching calibration – MMC

For MMC method, calibration curves in the range from 3.1 to 55 g kg<sup>-1</sup> Al and 0.72 to 11.6 g kg<sup>-1</sup> of Pb were obtained using four samples of waste PCBs as solid standards. The curves were obtained by plotting the analytical signal (y-axis emission intensity) *versus* the analyte concentration (x-axis).

Four emission lines with different relative intensities for AI (AI I 308.21 nm, AI I 309.40 nm, AI I 394.40 nm and AI I 396.15 nm) and two lines for Pb (Pb I 363.95 nm e Pb I 405.78 nm) were evaluated to obtain the calibration curves. The choice of the best normalization mode and the most appropriate emission line was made using as a criterion the obtaining of calibration curves that

enable the smallest errors of prediction of the analyte concentration in thesamples.

The concentration of the analyte, using MMC, is calculated using equation 1,

278 
$$C_{analyte} = \frac{Intensity-intercept}{slope}$$
 (Eq. 1)

where  $C_{analyte}$  is the concentration of AI or Pb determined in the sample, Intensity is the analytical signal of the emission line obtained, slope and intercept, both obtained by the calibration curve.

282 2.5.2. Two-point calibration transfer – TP CT

For TP CT only one sample is used as the calibration standard, and the linear model is obtained with two analytical signals monitoring only one analyte emission line. The linear model plot is made using two sets of spectra in the xaxis, and in the y-axis the intensity of the emission line is obtained using only the sum of the intensity of the spectra (height or signal area) through normalization 5.

Using the reference concentration ( $C_{standard}$ ) of the analyte in the calibration standard, and the slopes obtained in linear models for the sample (slope<sub>sample</sub>) and for the calibration standard (slope<sub>standard</sub>), the analyte concentration ( $C_{analyte}$ ) can be obtained using equation 2 (Castro et al. 2020).

293 
$$C_{analyte} = \frac{slope_{sample}}{slope_{standard}} \times C_{standard}$$
(Eq. 2)

For TP CT the emission lines AI I 396.15 nm and Pb I 405.78 nm were used to obtain linear models.

296 2.5.3. One-point and multi-line calibration – OP MLC

297 For the OP MLC, only one sample is used as the calibration standard 298 and several emission lines are used to obtain calibration linear models for Al

and Pb. In the x-axis, the emission intensities are used for all the monitored analyte lines in the standard calibration, and in the y-axis the intensities are used for all the monitored lines in the sample.

302 The analyte concentration ( $C_{analyte}$ ) is calculated using equation 3,

303

$$C_{analyte} = slope \times C_{standard}$$
(Eq. 3)

where the slope is obtained for the linear model, and  $C_{standard}$  is the concentration of the analyte in the sample used as the standard calibration (Hao et al. 2018).

Using four atomic emission lines for Al (308.21nm, 309.27 nm, 394.40 nm and 396.15 nm) and two atomic emission lines for Pb (363.95 nm and 405.78 nm), linear models were obtained for each analyte in the respective samples.

311 2.5.4. Single-sample calibration – SSC

In the SSC method, only one sample is used as the calibration standard and several emission lines of the analyte present in the standard and sample (unknown) are used. The SSC does not require a calibration curve or linear models (Yan et al. 2019).

For this strategy, the emission intensities of Al I 396.15 nm, Pb I 405.78 nm and Mg II 279.55 nm were used to determine the concentration of the analyte. The S2 PCB sample containing  $55 \pm 3$  g kg<sup>-1</sup> Al, 11.6  $\pm$  0.8 g kg<sup>-1</sup> Pb and 11.4  $\pm$  1.8 g kg<sup>-1</sup> Mg (obtained after microwave digestion and ICP OES determination) was used as the calibration standard.

321 The analyte concentration ( $C_{analyte}$ ) is calculated using equation 4,

322 
$$C_{analyte} = \frac{\frac{C \ standard \ analyte \times I \ analyte \ standard}{I \ analyte \ standard}}{\sum_{i=1}^{N} \frac{C \ standard \ element \times I^{N}_{element \ standard}}{I^{N}_{element \ standard}}}$$
(Eq. 4)

where  $C_{standard analyte}$  and  $I_{analyte standard}$  are the concentration and intensity of the emission line of the analyte in the PCB sample, respectively, used as the calibration standard (#S2 PCB). The  $I_{analyte sample}$  is the emission intensity of the analyte in the unknown sample. The  $I^{N}_{element sample}$  is the emission intensity of the element N in the sample of unknown concentration, and  $C^{N}_{standard element}$  and  $I^{N}$ element standard the concentration and the emission intensity of the element N, respectively, in the sample used as the standard calibration.

330

2.5.5. Calibration Free – CF

For CF only physicochemical parameters of the obtained plasma and from the monitored emission lines of the analytes and from all species present are necessary for the quantification. A calibration standard is not required.

334 The analyte concentration ( $C_{analyte}$ ) is calculated using equation 5,

335 
$$ln\frac{I_{\lambda}}{A_{ki}g_k} = ln\frac{C_{analyte}}{U(T)} - \frac{1}{K_BT}E_k$$
(Eq. 5)

where  $I_{\lambda}$  is the integrated intensity of the emission line,  $A_{ki}$  the transition probability,  $g_k$  the degeneration of the upper level,  $K_B$  the Boltzmann constant, T the temperature of the plasma,  $E_k$  the energy level of the excited state, F the experimental factor, and U(T) the partition function of the species present in the plasma (Ciucci et al. 1999, Li et al. 2019, Tognoni et al. 2010,).

All the calculations were processed using the LIBS++ software (ARWAN technology, developed by Palleschi et al.). For the calculation of the plasma temperature, the emission lines for Ba (Ba I 705.99, Ba II 614.17 and Ba II 649.69 nm), Fe (Fe I 374.55, Fe I 382.04, Fe I 405.58, Fe I 438.35, Fe II 239.92 nm) and Ti (Ti I 498.17, Ti I 499.10 and Ti II 333.94 nm) were used.

The plasma electron density ( $N_e$ ) was calculated from the H $\alpha$  line in 656.28 nm and using Equation 6,

348 
$$FWHA = 0.549nm \times \left(\frac{N_e}{10^{23} m^{-3}}\right)^{0.67965}$$
 (Eq. 6)

where FWHA denotes the full width at half area of this hydrogen emission line(Cavalcante et al. 2013, Ciucci et al. 1999,).

Figure 1 show a pictorial description of all the calibration strategies (calibration curve, linear model or correlation) used for Al.

353 2.6. Determination of analytical performance parameters

The precision (n=3) was calculated using all the samples. The standard error (SE) and root mean square error of prediction (RMSEP) were calculated for analytes, using Equations 7 and 8, respectively:

357 
$$SE = \sqrt{\frac{\sum(y_i - \hat{y})^2}{n-1}}$$
 (Eq. 7)

358 
$$RMSEP = \sqrt{\frac{\Sigma(y_i - \hat{y})^2}{n}}$$
(Eq. 8)

where  $y_i$  is the analyte reference concentration obtained by ICP OES,  $\hat{y}$  is the concentration predicted by the calibration model using LIBS, and n is the number of samples analyzed.

Slope and intercept values and respective confidence interval (95 % confidence level) obtained for linear regression for concentration reference (ICP OES method) *versus* concentration predicted (LIBS method) plots, were used for results comparison obtained in the direct determinations of analytes using five calibration strategies for LIBS. The ideal situation is a slope and intercept equals to 1 and 0, respectively.

368

#### 3. Results and discussion

369 3.1. Optmization of LIBS instrumental conditions

The instrumental conditions of the LIBS system influence the lasermatter interaction and also the quality of the emission spectrum obtained. The

laser pulse energy, delay time and gate width of the spectrometer were
optimized using a full factorial design 2<sup>3</sup> with center and axial points (see Table
S1).

The regression model based on the obtained OD (Table S1) was calculated to determine the best description of the experimental region. The quality of the model was evaluated through analysis of variance (ANOVA). After observing the values calculated for ANOVA, it was possible to verify that the regression of the model is not statistically significant at the 95% confidence level. These results demonstrate that it is not possible to obtain a model with good predictive capacity.

By evaluating Table S1, it was observed that experiment 8 presented the highest OD value (OD = 0.89) when compared to the other experiments. Thus, the evaluated conditions of this experiment were used in all measurements by LIBS in this study, with a delay time of 2  $\mu$ s, a gate width of 3  $\mu$ s and laser pulse energy of 42.5 mJ.

387 3.2. Laser-sample interaction: energy parameters

The physical and chemical properties of the sample strongly influence the laser-sample interaction and consequently the formation of the plasma, modifying its characteristics (temperature and electronic density, among others). Using SEM analysis and laser pulse energy optimized for analyses of waste PCBs, some parameters were obtained from the laser-waste PCB pellet interaction.

The crater formed by the laser pulse over the surface of the pelletized waste PCB sample S2 (200 mg compressed using  $10x10^4$  N for 2 min) is shown in Figure 2. The estimated crater diameter was 470 µm. The figure shows the

heterogeneity of the morphology and composition of the sample since waste PCBs consist of several polymeric, ceramic and metallic components. The importance of a milling step to obtain a representative sample and thus enable a stoichiometric ablation is also evident, besides the need to obtain several spectra in different regions of the pelletized sample for precision in determination (low RSD values). The laser-sample interaction is complex and many phenomena occur as a result.

The irradiance (W  $cm^{-2}$ ) and the laser pulse fluence (J  $cm^{-2}$ ) were 404 calculated from the diameter of the crater. Using a laser pulse of 42.5 mJ and a 405 pulse duration of 7.7 ns, a power of 5.5 MW was generated. Even if low 406 energies are used, it is common to obtain high power values because the pulse 407 duration lasts for nanoseconds. The crater radius reached 235 µm, obtaining an 408 irradiance of 3.2 GW cm<sup>-2</sup> and a laser fluence of 24.5 J cm<sup>-2</sup>. These parameters 409 were calculated for the optimized instrumental conditions used to obtain all the 410 411 LIBS spectra for the waste PBC samples.

#### 412 3.3. Evaluation of calibration strategies for LIBS

The matrix effects are the main sources of the linearity deviations 413 between concentration and emission intensity in the analysis of solids by LIBS 414 aimed at elementary determination. Thus, five calibration strategies (i- MMC, ii-415 TP CT, iii- OP MLC, iv- SSC and v- CF) were evaluated to overcome or 416 minimize matrix effects in the determination of AI and Pb in six waste PCBs by 417 LIBS. The criterion for selecting the analytes emission lines ( $\lambda$ ) used in each 418 calibration strategy was made considering the accuracy of the determinations 419 based on the recovery value. Recoveries values in the range of 80 to 120% 420 were considered satisfactory for all the evaluated calibration strategies. 421

Using matrix-matching calibration for AI, a calibration curve (coefficient 422 of determination:  $R^2$  = 0.8146) was obtained by monitoring the emission line AI I 423 396.15 nm in the following samples: S1, S2, S3 and S4. For the validation of the 424 method, two samples with intermediate concentrations to the calibration 425 standards (S5 and S6) were analyzed. Recoveries values of 99% and 116%, 426 standard deviation (RSD) values  $\leq 4\%$  were obtained, and relative 427 demonstrating the satisfactory accuracy of MMC for AI determinations (see 428 429 Table 1).

For Pb, the calibration curve obtained monitoring the emission line Pb I 431 405.78 nm at samples S2, S3, S4 and S5, showed a good coefficient of 432 determination ( $R^2$ = 0.8426) using the MMC strategy. For samples S1 and S6 433 (used for validation), good recoveries values of 102% and 111% and RSD  $\leq$ 434 8% were obtained for Pb determinations by LIBS (see Table 2).

For two-point calibration transfer strategy, the S4 (13.4  $\pm$  0.7 g kg<sup>-1</sup> Al) 435 and S1 (7.6  $\pm$  0.6 g kg<sup>-1</sup>Pb) samples were used as calibration standards for AI 436 and Pb, respectively. Since only two calibration points ("concentrations") are 437 used, the linearity and significance of the model can be verified from the test F, 438 and in this case the ratio  $F_{experimental}/F_{tabulated}$  was calculated. This ratio  $\geq 10$ 439 demonstrated that the variances are statistically different (the quadratic mean of 440 441 the regression is statistically different when compared with the quadratic mean of the residues), thus the model can be considered linear and statistically 442 significative, and the TP CT can be used (Pereira and Pereira-Filho 2018). The 443 ratio found for all samples analyzed for AI ranged from 7 to 416 and for Pb 444 ratios of 5 to 116 were obtained, indicating that the models are linear and that 445 two-point calibration is feasible. 446

For Al, recovery values ranging from 84 to 112% were obtained using 447 TP CT, except for S1 (683%), S2 (125%) and S3 (61%). For Pb, recovery 448 values ranging from 89 to 104% were obtained, except for sample S4 (132%) 449 (Tables 1 and 2). It was not possible to determine Pb in the S3 sample using 450 LIBS and TP CT (or any other calibration strategy evaluated), since the 451 concentration in this sample  $(0.72 \pm 0.09 \text{ g kg}^{-1}\text{Pb})$  is lower than the standard 452 error (SE) calculated for the LIBS method (Table 3). RSD values ≦17% and 453  $\leq$ 22% were obtained in the determinations of AI and Pb, respectively, using TP 454 CT. 455

For one-point and multi-line calibration, the samples of waste PCBs used as solid standards for AI and Pb were S5 (10.2  $\pm$  1 g kg<sup>-1</sup> AI) and S2 (11.6  $\pm$  0.8 g kg<sup>-1</sup> Pb), respectively. Good linear models were obtained, with excellent coefficients of determination for AI (R<sup>2</sup> ranging from 0.9790 to 0.9960) and Pb (R<sup>2</sup>= 1), using four atomic emission lines for AI and two lines for Pb.

For AI, recoveries values ranging from 78 to 109% were obtained using 461 OP MLC, except for S3 (57%). For Pb, excellent recoveries values ranging from 462 83 to 103% were obtained for all the samples analyzed (Tables 1 and 2). The 463 values of the experimental slopes calculated for the linear models do not 464 present significant differences from the theoretical slope (see Tables 1 and 2), 465 providing good values of recoveries and demonstrating a satisfactory accuracy 466 of the determinations. RSD values  $\leq 9\%$  and  $\leq 25\%$  were obtained in the 467 determinations of AI and Pb, respectively. 468

Using the single-sample calibration for AI determination, the emission lines and concentrations of AI and Pb for all the samples and standard were used, except for samples S1 and S3. For these two samples, the intensity of the

emission lines of Al, Pb and Mg and the respective concentrations of these
elements in the calibration standard were used. Recoveries values ranging from
82 to 116% were obtained using SSC, except for S1 (220%).

475 However, for Pb determination the emission intensity of the Pb, Al and Mg lines was used for all the waste PCB samples, together with the respective 476 concentration values of these three elements in the calibration standard, except 477 for sample S1. Only the Pb and Mg elements were monitored for this sample, 478 together with their respective concentrations in the standard for correlation with 479 the S1 sample. Recoveries values ranging from 81 to 116% were obtained 480 481 using SSC, except for S6 (71%) - see Table 2. Using the SSC as a calibration 482 strategy, RSD values of  $\leq 21\%$  and  $\leq 25\%$  were obtained in the determinations of AI and Pb, respectively. 483

Calibration free was another strategy evaluated for the determination of the analytes in the complex and refractory waste PCB samples. For CF it is necessary to obtain some physical parameters of the plasma, such as the temperature and electronic density, to verify the local thermodynamic equilibrium (LTE) (Ciucci et al. 1999, Tognoni et al. 2010).

For the calculation of the plasma temperature using the Saha-Boltzmann equation, the emission intensities of different lines in different ionization states (atomic and ionic) for Ba, Fe and Ti were used. The average plasma temperature was  $8145 \pm 227$  K, considering the six waste PCB samples analyzed. The physical parameters of all the elements evaluated in the CF-LIBS are shown in Table S2, see supplementary material.

495 The electron density was calculated from the collision-induced 496 enlargement of the Balmer  $H_{\alpha}$  line to the hydrogen. The average plasma

electron density was  $0.65 \pm 0.29 \ 10^{17} \text{ cm}^{-3}$ , considering all the samples analyzed.

From these values obtained for the temperature and electronic density of the plasma (McWhirter criterion) (Tognoni et al. 2010) it is possible to assure the existence of LTE in all the samples analyzed. Thus, also taking into account stoichiometric ablation and using optically thin plasma, the concentration of Al and Pb in the samples can be determined.

For AI, excellent recovery values ranging from 90 to 106% were obtained using CF for all the samples analyzed. The recovery values for Pb ranged from 78 to 121%, except for S6 (162%) - see Tables 2 and 3. RSD values of  $\leq$ 26% were obtained in AI and Pb determinations.

508 Some parameters related to the analytical performance and processing 509 of the data used for each of the evaluated calibration strategies are shown in 510 Table 3. It is interesting to observe how these parameters and the processing of 511 the spectra can change depending on the calibration strategy (SE and RMSEP, 512 for example). This is an indication of how matrix effects can be minimized by 513 using appropriate data processing and calibration strategies.

514 From the equations of the linear regression of the validation set (ICP 515 OES concentration reference *versus* LIBS predicted concentration plot), it is 516 possible to see that, in almost all cases the values, considering the confidence 517 interval, that the for the angular coefficient interval includes the number 1 and 518 the intercept includes the number 0, see Figures S1 and S2 in supplementary 519 material.

3.4. What is the best calibration strategy for the determination of Al andPb in waste PCBs by LIBS?

In the plasma induced by LIBS, physicochemical phenomena and 522 523 matrix effects occur during and due to laser-sample interaction, which in turn act on the atomic emission of the analytes, directly influencing the determination of 524 525 Al and Pb in the samples of waste PCBs. However, with the data obtained from the evaluated calibration strategies, it can be seen that some of these strategies 526 527 were very efficient, producing results with satisfactory accuracy. It can also be seen that the intrinsic properties of each analyte and of each calibration strategy 528 directly influence the choice of the best calibration strategy. 529

The matrix-matching calibration gave excellent results for the determination of both analytes, with satisfactory recovery and RSD values. The MMC proved to be an efficient calibration strategy for the analysis of solids by LIBS, because since a set of waste PCB samples were used as calibration standards, the possible matrix effects are minimized when the physical properties of the calibration standards are close to those of the analyzed samples.

One limitation of the use of MMC in this study is that there is no set of 537 certified reference materials (CRM) of waste PCBs, with reference values for AI 538 and Pb, which could be used as solid standards when obtaining the calibration 539 curve. Few initiatives are observed in the literature in order to produce a 540 reference material for WEEE and a good example was published by Andrade et 541 al. 2019a, 2019c. Thus, it was necessary to first obtain reference values of the 542 543 analytes by analyzing a set of samples by a reference technique (in this study an ICP OES was used) for use as calibration standards for the LIBS method. 544

545 In some cases it is necessary to use vigorous conditions for the 546 decomposition of the samples (high temperatures and high volume of

547 concentrated acid, as in this study). This represents a limitation in the use of 548 MMC in the absence of adequate sample preparation instruments and 549 reference values of the analytes required for later use of the samples as solid 550 calibration standards for LIBS.

Two-point calibration transfer requires only one calibration standard (CRM or one reference value sample) and one sample with unknown concentration to determine the analyte concentration. In this calibration strategy, each set of spectra obtained in the LIBS analysis for standard and sample are divided into two sets and subsequently summed (the number of spectra composing set 2 must have approximately two-fold the number of spectra of set 1) (Castro et al. 2020).

If the standard and the sample have similar physical properties (for 558 559 efficient matrix-matching), and the concentration of the standard is close to that of the sample, TP CT minimizes the matrix effects and enables a high degree of 560 561 accuracy when determining the analyte concentration in the sample, using only 562 one calibration standard and one linear model with two points. Tables 1 and 2 show that good recoveries values (ranging from 80 to 120%) were obtained for 563 Al and Pb, when the concentration of the standard used was close to the 564 concentration of these analytes in the sample. For samples with concentrations 565 very different from those of the standards used, there was an under- or over-566 estimation of the analyte concentration. 567

The TP CT is an interesting simple calibration strategy for LIBS analysis when there is not a great variability of analyte concentration in the analyzed samples and when a standard with an appropriate concentration similar to that of the samples is used. This situation can be achieved in routine analysis.

Another strategy evaluated was one-point and multi-line calibration - OP 572 MLC. Excellent results were obtained for all the determinations of AI (except for 573 sample S3) and Pb in the waste PCB samples by LIBS. The OP MLC requires 574 575 only one calibration standard and several lines of analyte emission to obtain the linear model, which facilitates the implementation of this calibration strategy 576 when few solid calibration standards are available in the laboratory routine. 577 Extra care has to be taken when using the OP MLC to remove lines with low 578 intensity that present spectral interferences, since they can harm the linear 579 models and consequently the measurement accuracy (Hao et al. 2018). 580

The determination of AI and Pb was also evaluated using the singlesample calibration method - SSC, which is another recent calibration strategy for LIBS that uses only one sample as standard (reference). In this strategy, a simple correlation calculation is necessary to determine the analyte concentration in the sample with unknown concentration (Yan et al. 2019). Using SSC, good recoveries values were obtained for both AI (except sample S1) and Pb (except sample S6).

For the use of SSC, the extent of the matrix effects between the sample 588 and the standard should be considered the same for all elements present in the 589 LIBS-induced plasma sample and the standard, since a direct correlation 590 between the emission intensity and concentration of these elements will be 591 used for the determination of the analyte. In addition, it should be considered 592 that none of the emission lines used in the SSC (analyte emission lines and 593 other elements used in the correlation) present spectral interferences, so that 594 results can be obtained with satisfactory precision and recovery (Yan et al. 595 2019). 596

597 Finally, using calibration free (CF) it was possible to obtain good results 598 with excellent accuracy for AI and Pb determinations, except for AI 599 determination in sample S6, using direct analysis of waste PCB samples by 600 LIBS.

Not requiring a calibration standard is an advantage of CF, which 601 makes it an excellent calibration strategy for use in elemental determination by 602 LIBS in complex samples difficult to decompose. However, the quality of data 603 604 acquisition and treatment in CF is a critical factor for obtaining satisfactory results. It is necessary to ensure that the emission measurements are obtained 605 606 in LTE in the plasma, that the physical parameters used are correctly obtained and calculated with precision, and that only emission lines free of spectral 607 interferences and self-absorption are employed in CF. Despite it is a laborious 608 609 calculation procedure, CF allows the achievement of good results.

In this context, it is evident that it is difficult to choose the best calibration strategy for the direct determination of AI and Pb in waste PCBs by LIBS, since it depends on many variables. However, a knowledge of the advantages and limitations of each calibration strategy and a consideration of some intrinsic characteristics (physicochemical properties) of the sample and the analytes can help in selecting the best strategy that efficiently overcomes matrix effects and enables determination with satisfactory accuracy.

Table 4 shows some characteristics, advantages and limitations of the five calibration strategies evaluated in this study. This may help the reader to choose and evaluate the best calibration strategy for LIBS that could be used in different analytical contexts.

It is worth noting that there seems to be a tendency in recently reported 621 new calibration strategies for LIBS to use no or few calibration standards (only 1 622 or 2) and to increasingly explore the physicochemical parameters and 623 correlations of concentrations of the species present in the plasma induced by 624 laser in each sample analyzed. Examples are TP CT, OP MLC, SSC evaluated 625 in this study and other strategies recently reported in the scientific literature 626 such as MEC (Babos et al. 2018) and OP GSA (Babos et al. 2019, Castro et al. 627 2020) calibration. 628

3.5 Evaluation of AI and Pb concentrations in waste PCBs: economicand environmental questions

As mentioned previously, waste PCBs can contain high concentrations of valuable and toxic metals. The recycling and appropriate disposal of this waste can both provide a source of income and contribute to environmental protection. In this context, LIBS is an excellent analytical tool for the monitoring of metals in the waste and for the development of methods for the analysis of solids and the direct determination of Al and Pb present in PCBs.

Were calculated estimations of the commercial value (in US\$) per ton of each of the six samples of waste PCBs analyzed, considering only the measured concentrations of AI (ranging from 3.1 to 55 g kg<sup>-1</sup> AI) and Pb (ranging from 0.72 to 11.6 g kg<sup>-1</sup> Pb). Considered the prices of 1803 US\$/tonne AI and 1919 US\$/tonne Pb quoted on the London Metal Exchange, the second world center for industrial metals trading (LME, 2019).

The commercial value per tonne of the analyzed samples, considering only the AI and Pb contents, range from 20 to 121 US\$/tonne of waste PCB (prices for the S1 and S2 samples, respectively). These are good prices for

samples that are considered waste, particularly as they apply only to two metals
present in the waste. Other valuable metals may also be present and thus the
market price per ton of waste PCBs may be higher.

649 Concerning the environmental question, only one sample (S3 PCB, 0.072% Pb) complies with the maximum concentration value allowed (0.1% Pb) 650 by weight in homogeneous materials for Pb in WEEE under Directive 651 2011/65/EU (RoHS, 2011). The Pb concentration in the other samples is 652 between 5 and 12 times above the maximum allowed value according to the 653 normative instruction. This is worrying, since if these samples are 654 inappropriately disposed of they may be a source of contamination in the 655 environment because of the Pb metal content. 656

657

#### 658 **Conclusion**

The choice of the best calibration strategy for the direct analysis of waste PCBs 659 660 for LIBS when aiming to determine AI and Pb depends on the intrinsic properties of these analytes and samples, as well as the ability of each of the 661 calibration strategies to overcome the various matrix effects. Of the five 662 calibration strategies evaluated, MMC and CF generally allowed accurate 663 values to be obtained for both analytes in all the samples. The LIBS technique 664 presented itself as an excellent analytical tool in the fast, simple and direct 665 monitoring of recyclable metals such as AI and Pb and also of potential for 666 environmental contamination such as Pb, originating from WEEE (waste PCB). 667 The Pb concentrations determined are of concern as only one sample was in 668 accordance with the Directive 2011/65/EU. 669

670

#### 671 **Credit author statement**

Diego Victor Babos: Conceptualization; Data curation; Formal analysis;
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### 680 **Declaration of Competing Interest**

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

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# **Table captions**

**Table 1.** Concentrations (mean ± standard deviation, g kg<sup>-1</sup> Al, n=3) and recovery (%) for Al in PCB samples determined by LIBS using matrix-matching calibration (MMC), two-point calibration transfer (TP CT), one-point and multi-line calibration (OP MLC), single-sample calibration (SSC) and calibration free (CF).

	ICP OES	MMC	TP CT		OP MLC		SSC	CF
Sample	g kg <sup>-1</sup>			Theoretical slope*	Theoretical Experimental slope* Slope			
S1	3.1 ± 0.4	-	21.2 ± 3.6 (683)	0.30	0.28 ± 0.03	2.9 ± 0.3 (93)	6.8 ± 1.1 (220)	3.3 ± 0.1 (106)
S2	55 ± 3	-	68.9 ± 8.4 (125)	5.39	5.87 ± 0.11	59.8 ± 1 (109)	-	49.8 ± 9.2 (91)
S3	21.4 ± 0.6	-	13.0 ± 1.4 (61)	2.09	1.19 ± 0.07	12.1 ± 0.6 (57)	24.8 ± 5.1 (116)	21.4 ± 3.1 (100)
S4	13.4 ± 0.7	-	-	1.32	1.02 ± 0.08	10.4 ± 0.8 (78)	13.5 ± 1.4 (101)	13.1± 3.4 (98)
S5	10.2 ± 1	11.8 ± 0.1 (116)	11.4 ± 0.9 (112)	-	-	-	10.4 ± 2.1 (102)	10.0 ± 2 (98)
S6	15.2 ± 0.9	15.0 ± 0.5 (99)	12.7 ± 1.4 (84)	1.49	1.23 ± 0.05	12.6 ± 0.5 (83)	12.4 ± 0.6 (82)	13.6 ± 1.2 (90)

- : Samples used for calibration

\*Theoretical slope:  $slope = \frac{C_{analyte}}{C_{standard}}$ 

**Table 2.** Concentrations (mean ± standard deviation, g kg<sup>-1</sup> Pb, n=3) and recovery (%) for Pb in PCB samples determined by LIBS using matrix-matching calibration (MMC), two-point calibration transfer (TP CT), one-point and multi-line calibration (OP MLC), single-sample calibration (SSC) and calibration free (CF).

	ICP OES	MMC	TP CT		OP MLC		SSC	CF
Sample		g kg⁻¹		Theoretical slope*	Experimental Slope		g kg⁻¹	
S1	7.6 ± 0.6	7.8 ± 0.5 (102)	-	0.655	0.574 ± 0.089	6.7 ± 1 (88)	7.3 ± 0.6 (96)	5.9 ± 0.6 (78)
S2	11.6 ± 0.8	-	11.6 ± 0.8 (100)	-	-	-	-	14.1 ± 3.6 (121)
S3	$0.72 \pm 0.09$	-	<1.1	0.062	$0.008 \pm 0.002$	<1.2	<1.7	<2.7
S4	4.7 ± 0.9	-	6.2 ± 0.1 (132)	0.405	0.413 ± 0.029	4.8 ± 0.3 (103)	5.4 ± 0.3 (116)	5.5 ± 0.8 (116)
S5	10.7 ± 0.6	-	9.5 ± 2.1 (89)	0.922	0.766 ± 0.059	8.9 ± 0.7 (83)	8.6 ± 2.1 (81)	10.3 ± 2.2 (97)
S6	6.9 ± 1.6	7.6 ± 0.6 (111)	7.2 ± 0.9 (104)	0.595	0.611 ± 0.160	7.1 ± 1.8 (103)	4.9 ± 0.8 (71)	11.2 ± 2.2 (162)

- : Samples used for calibration

\*Theoretical slope:  $slope = \frac{C_{analyte}}{C_{standard}}$ 

 Table 3. Analytical performance parameters

Parameter	Matrix-matching	Two-point	Calibration	One point and multi-	line Single-sample
	calibration	calibration	free	calibration	calibration
		transfer			
Emission line	AI I 396.15	AI I 396.15	Several lines*	Al I 308.21 Pb I 363	.95 AI I 396.15
(nm)	Pb I 405.78	Pb I 405.78		AII 309.27 Pb I 405	.78 Pb I 405.78
				AI I 394.40	Mg II 279.55
				AI I 396.15	
Normalization selected					
Al	Each individual spectrum	Sum	Average of the	Each individual spectrun	n is Each individual spectrum is
	is divided by its Euclidean		spectra	divided by its Euclidean n	orm, divided by its Euclidean
	norm and the average is			and the average is calculate	ated norm and the average is
	calculated				calculated
Pb	Average of the spectra	Sum	Average of the	Average of the spectra	Average of the spectra
			spectra		
Signal type					
Al	Height	Height	Area	Height	Area
Pb	Height	Height	Area	Area	Height
SE (g kg⁻¹)					
Al	1.6	12.2	2.4	5.6	2.9
Pb	0.76	1.1	2.7	1.2	1.7
RMSEP (g kg <sup>-1</sup> )					
AI	1.1	10.9	2.2	5.0	2.6
Pb	0.49	0.87	2.4	0.91	1.3
RSD range (%)					
AI	1 - 4	8 - 17	3 - 26	2 - 9	5 - 21
Pb	6 - 8	2 - 22	10 - 26	6 - 25	6 - 25

\* see Table S2 in Electronic Supplementary Information (ESI).

**Table 4.** Advantages and limitations of the calibration strategies evaluated in this study for the determination of AI and Pb in waste PCBs by LIBS.

Calibration Method	Number of emission lines	Number of reference standard	Advantages	Limitations	Reference	
Matrix- matching calibration	One	Between four and five	Efficient matching of the physical properties of the calibration standards with the samples	Difficulty in obtaining a set of commercial CRM or samples with reference values	Vieira et al. 2018, this study	
Calibration free	Several	Not necessary	Does not require the use of calibration curves or matrix-matched standards	Occurrences of lines with self- absorption and spectral interferences, which compromise the determinations, laborious calculations	Tognoni et al. 2010, this study	
One point and multi-line calibration	Several	One	Requires only one calibration standard	The choice of the standard with appropriate concentration, and the use of emission lines with low sensitivities	Hao et al. 2018, this study	
Two-point calibration transfer	One	One	Simplicity in data processing and measurement accuracy	The choice of the standard with appropriate concentration as samples, and standards should show little variability between analyte concentrations	Castro et al. 2020, this study	
Single- sample calibration	The number of lines of analytes present in the sample shall be ≤ to the number of lines used of the elements in the standard	One	No calibration curve or linear calibration model required	Samples and standards with significant variability of physico- chemical properties and analyte emission lines and reference elements with spectral interferences	Yan et al. 2019, this study	

# **Figure captions**

**Figure 1.** Scheme representation of the calibration strategies evaluated for direct determination of AI in waste PCBs by LIBS.

**Figure 2.** Crater and superficial morphology of waste PCB sample (S2) obtained by scanning electron microscopy (SEM).



Figure 1





**Electronic Supplementary Information (ESI):** 

Direct determination of AI and Pb in waste printed circuit boards (PCB) by Laser-induced breakdown spectroscopy (LIBS): evaluation of calibration strategies and economic - environmental questions

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This ESI contains:

- 2 Tables
- 2 Figures

Experiment		Delay time (µs)		Gate w	Gate width (µs)		Laser pulse energy (mJ)	
		Real	Coded	Real	Coded	Real	Coded	
33	1	0.4	-1	1.0	-1	25	-1	0.49
uf	2	2.0	1	1.0	-1	25	-1	0.73
esic	3	0.4	-1	3.0	1	25	-1	0.67
l de	4	2.0	1	3.0	1	25	-1	0.55
oria	5	0.4	-1	1.0	-1	42.5	1	0.57
acto	6	2.0	1	1.0	-1	42.5	1	0.62
	7	0.4	-1	3.0	1	42.5	1	0.30
Ц Ц	8	2.0	1	3.0	1	42.5	1	0.89
t	9	1.2	0	2.0	0	35	0	0.44
poi	10	1.2	0	2.0	0	35	0	0.56
al	11	1.2	0	2.0	0	35	0	0.77
enti	12	1.2	0	2.0	0	35	0	0.61
ŏ	13	1.2	0	2.0	0	35	0	0.62
	14	0.01	-1.49	2.0	0	35	0	0.41
ц	15	2.54	1.68	2.0	0	35	0	0.74
poi	16	1.2	0	0.32	-1.68	35	0	0.63
la	17	1.2	0	3.68	1.68	35	0	0.64
Ax	18	1.2	0	2.0	0	20	-1.68	0.69
	19	1.2	0	2.0	0	47.5	1.68	0.69

**Table S1.** Matrix of experiments showing the variables evaluated for optimizing delay time, gate width and laser pulse energy in LIBS determinations. The overall desirability (OD) was used as experimental response.

**Table S2.** Spectroscopic parameters  $\lambda$  (wavelength),  $E_i$  (energy of the lower level of transition),  $E_k$  (energy of the upper level of transition),  $A_{ki}$  (transition probability), and  $g_k$  (degeneracy factor of state k) of atomic (I) and ionic (II) lines used in the CF-LIBS calculation. Source: NIST DataBase.

Elements	Line	λ <b>(nm)</b>	E <sub>i</sub> (eV)	E <sub>k</sub> (eV)	Яĸ	A <sub>ki</sub> .10 <sup>8</sup> (s <sup>-1</sup> )
AI	I	309.271	0.01	4.02	6	0.730
AI	I	783.531	4.02	5.60	6	0.057
AI	I	783.613	4.02	5.60	6	0.004
Ва	II	614.172	0.70	2.72	4	0.412
Ва	II	649.690	0.60	2.51	2	0.130
Ba	I	705.994	1.19	2.95	9	0.500
С	I	247.856	2.68	7.68	3	0.180
Ca	II	317.933	3.03	7.05	6	3.600
Ca	I	558.876	40.05	43.51	7	0.409
Ca	I	612.222	1.89	3.91	3	0.287
Ca	I	616.217	1.90	3.91	3	0.477
Co	I	356.938	0.92	4.40	8	1.500
Cr	II	283.563	1.55	5.92	12	2.000
Cr	I	360.533	0.00	3.44	5	1.620
Cu	I	510.554	1.39	3.82	4	0.020
Cu	I	521.820	3.82	6.19	6	1.220
Cu	I	578.213	1.64	3.79	2	0.019
Fe	II	239.924	0.08	5.25	6	1.400
Fe	I	373.532	0.86	4.18	7	0.270
Fe	I	382.042	0.86	4.10	9	0.668
Fe	I	404.581	1.48	4.55	9	0.863
Fe	I	438.354	1.48	4.31	11	0.500
Mg	I	518.361	2.72	5.11	3	0.561
Mn	II	293.306	1.17	5.40	3	2.040
Ni	I	341.476	0.03	3.66	9	0.550
Ni	I	351.505	0.11	3.64	7	0.420
Ni	I	352.454	0.03	3.54	5	1.000
Pb	I	405.780	1.32	4.38	3	0.912
Sb	I	259.805	1.06	5.83	2	0.210
Si	I	288.158	0.78	5.08	3	1.890
Sn	I	317.050	0.42	4.33	3	0.838
Ti	II	334.941	0.05	3.75	12	1.680
Ti	I	498.173	0.85	3.34	13	0.660
Ti	I	499.107	0.84	3.34	11	0.584
Zn		481.053	4.08	6.65	3	0.700

**Figure S1.** Comparison of AI concentrations determined in waste PCBs samples by the proposed LIBS method using different calibration strategies (a-MMC, b- TP CT, c- OP MLC, d- SSC and e- CF) and the ICP OES reference method. The SE and RMSEP were added as lines parallel to the X-axis.







**Figure S2.** Comparison of Pb concentrations determined in waste PCBs samples by the proposed LIBS method using different calibration strategies (a-MMC, b- TP CT, c- OP MLC, d- SSC and e- CF) and the ICP OES reference method. The SE and RMSEP were added as lines parallel to the X-axis.





