# 1 Burn effects on soil properties associated to heat transfer under

# 2 contrasting moisture content

- 3 David Badía<sup>\*</sup>, Sergio López, Clara Martí, Oriol Ortíz, Antonio Girona, José 4 5 Casanova 6 7 <sup>a</sup>Departamento de Ciencias Agrarias y del Medio Natural. Escuela Politécnica Superior, University of 8 Zaragoza, Crtra. Cuarte s/n, Huesca 22071, Spain. 9 \*Corresponding author: badia@unizar.es 10 11 Keywords: fire, water content, soil depth, water repellency, organic C, EC 12 13 HIGHLIGHTS 14 High soil moisture content attenuates the heat transmission within the mollic horizon 15 Soil thickness affected by burning is mainly limited to the first cm, either wet or dry 16 17 Fire leads to a C loss of 50% in dry soil and 25% in wet soil for up to the 1<sup>st</sup> cm 18 SWR is tightly linked with fire, soil moisture content and soil depth 19 20 21 ABSTRACT 22 The aim of this work is to investigate the topsoil thickness affected by burning under contrasting soil moisture content (field capacity versus air-dried conditions). A mollic horizon 23 of an Aleppo pine forest was sampled and burned in the laboratory, recording the temperature 24 continuously at the topsoil surface and at soil depths of 1, 2, and 3 cm. Changes in soil 25 26 properties were measured at 0-1, 1-2, 2-3, and 3-4 cm. Both the maximum temperature and the charring intensities were significantly lower in 27 28 wet soils than in air-dried soils up to 3 cm in depth. Moreover, soil heating was slower and cooling faster in wet soils as compared to dry soils. Fire caused an immediate and significant 29 decrease in water repellency in the air-dried soil, even at 3 cm depth, whereas the wet soil 30 remained hydrophilic throughout its thickness, without being affected by burning. Burning 31
- 32 depleted 50% of the soil organic C (OC) content in the air-dried soil and 25% in the wet soil

at the upper centimeter, which was blackened. Burning significantly decreased the total N 1 (TN) content only in the dry soil (to one-third of the original value) through the first 2 centimeter of soil depth. Soluble ions, measured by electrical conductivity (EC), increased 3 after burning, although only significantly in the first centimeter of air-dried soils. Below 2 cm, 4 burning had no significant effects on the color, OC, TN, or EC, for either wet or dry soil. 5 Meanwhile, the water was vaporized and the soil absorbed a certain amount of heat without a 6 7 subsequent temperature increase, which attenuated heating effects on soil properties. The 8 effect of fire was limited to the upper centimeters, either wet or dry.

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## 10 **1. Introduction**

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Wildfires may induce changes in many soil properties and components, such as soil 12 13 organic matter (DeBano et al., 1998; Certini et al., 2011; Almendros and González-Vila, 2012), nutrient availability (Giovannini, 2012; Badía et al., 2014a), or soil water repellency 14 (Badía et al., 2013a; Bodí et al., 2013; Keesstra et al., 2017). These changes, combined with 15 16 the removal of plant cover after burning, may lead to direct or indirect soil loss (Badía and Martí, 2008; Shakesby, 2011; Vega et al., 2014; García-Orenes et al., 2017). The maximum 17 temperature reached at the mineral soil surface during a fire has been positively related to the 18 19 accumulated soil loss during the first year after the fire (Fernández et al., 2008; Vega et al., 2014). The impact of fire reported in the literature varies from negligible to very severe 20 21 depending on several thresholds and factors (Santín and Doerr, 2016), e.g., the natural variability between the study sites in terms of soil moisture content (Campbell et al., 1995; 22 23 Vadilonga et al., 2008; Santín et al., 2016), soil type (Badía and Martí, 2003a), and vegetation type (Bento-Gonçalves et al., 2012; Cawson et al., 2012). Additionally, some discrepancies in 24 25 the effects of a fire on soil properties can result from the sampling time (Pereira et al., 2012) 26 and even the scale of the measurement (Cawson et al., 2016). Another relevant source of 27 variability is related to fire characteristics or thermal signature (maximum temperature, heat 28 duration, and oxygen availability) and heterogeneous spatial heating, even in experimental 29 burns (Santín et al., 2016; Vega et al., 2014).

Other discrepancies described in the literature about the effects of fire on soil have been related to the soil depth sampled and the mixture of ashes and charred organic material together with the mineral topsoil (Certini et al., 2011). Because soil is a poor conductor of heat (DeBano, 2000), there is a strong gradient of temperatures with soil depth during

wildfires and prescribed fires (Úbeda and Outeiro, 2009; Cawson et al., 2012). For these 1 reasons, direct fire-induced changes in soils are mainly noticeable in the organic layer or in 2 the most superficial centimeters of the mineral topsoil (Humphreys and Craig, 1981; 3 Fernández et al., 2013; Vega et al., 2014). Nevertheless, high-intensity burns, e.g., with a 4 5 large fuel load, affect mineral topsoil at the decimetric (up to 15 cm) level (Ulery et al., 2017). An accurate sampling, centimeter by centimeter, is required to acknowledge the affected soil 6 7 thickness and to avoid dilution effects (Badía et al., 2013b, 2016; San Emeterio et al., 2016). However, this detailed soil sampling is subjected to several problems in the field, such a high 8 9 spatial variability (Cawson et al., 2016), thus generating wide variations between replicates and making the interpretation of results difficult. 10

In addition, heat transmission into soil can be modified under high moisture and low 11 12 temperature conditions. For instance, Santín et al. (2016) considered that the experimental burn of boreal forest has negligible direct effects on the moist mineral soil. After prescribed 13 burning of the understory in Catalonia (NE Spain) forests, Vadilonga et al. (2008) observed 14 that the presence of moist organic litter prevented the underlying soil from excessive heating. 15 The use of prescribed fire to control shrub encroachment in grasslands in European mountains 16 is recommended in wet seasons (De Partearroyo et al., 2012). However, recent studies on this 17 subject have had opposite results (Armas-Herrera et al., 2016; San Emeterio et al., 2016; 18 Girona et al., 2017). As DeBano et al. (1998) warned, heat transfer in wet soils is complex 19 20 because water increases both the heat capacity and the heat conductivity of the soil, properties 21 with opposite effects.

Predicting changes in soil properties induced by fires in the field often yields inconsistent results due to the soil thickness sampled, the mixture of ashes and charred material with the mineral soil, the sampling time, the soil moisture content, and the soil type. Moreover, heat transfer behavior is not always well known, and, therefore, burning soils under controlled conditions could be a useful method to check specific changes, centimeter by centimeter, on soil properties (Badía et al., 2013a, 2013b, 2014a, 2014b; Prat-Guitart et al., 2016).

The objective of this research was to examine the burn effect, under contrasting soil moisture content conditions (air-dried versus field conditions), on the heat transfer and several key physical and chemical soil properties of a mollic horizon at the centimeter scale (Ah horizon at depths of 1, 2, 3, and 4 cm).

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## 34 **2. Material and methods**

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# 2.1. Soil description and soil sampling

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4 Topsoil was sampled from a long-unburned, wooded, mountainous environment in the 5 northern foothills of Montes de Zuera (NE Spain), 630 m above sea level. This topsoil is a 6 mollic horizon from a Rendzic Phaeozem, under a dense Aleppo pine forest (*Pinus* 7 *halepensis*) with kermes evergreen oak (*Quercus coccifera*) and Mediterranean false-brome 8 (*Brachypodium retusum*). Dry topsoil is strongly water repellent, with a clay loamy texture 9 (Table 1). Additional details of the soils and the study area have been previously published 10 (Badía et al., 2013b).

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**Table 1.** Main physical and chemical properties of the experimental mollic Ah horizon (0–30 cm).

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To keep the original soil structure, nine unaltered topsoil monoliths were carefully removed with a steel cylinder (15 cm diameter x 6.5 cm deep) and transported to the laboratory. Although it is common to find an O-Ah sequence, the O horizon was removed before burning since the role of this layer was previously addressed in a similar study (Badía et al., 2014b). Six monoliths were air-dried to constant weight in the laboratory. Then, three of them were selected as controls (unburned) and the other three were burned (burned dry); additionally, three more monoliths were moistened to field capacity and then burned (burned wet).

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22 2.2. Burning treatment and characterization of the fire's thermal signature

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Each soil monolith was burned separately in a combustion tunnel by applying a flame 24 with a blowtorch placed 0.4 m above. The heating time was 15 min and the temperature was 25 26 continuously recorded with four type-K thermocouples (1 mm in diameter) placed at the soil surface and at depths of 1, 2, and 3 cm. The thermocouples were connected to a datalogger 27 (Picotech TC-08), and data were recorded every 8 seconds until cooling (~210 min), using the 28 29 PicoLog R5.21.9 program. The heating duration for a range of temperatures, the maximum temperature (Tmax), and the charred intensity (CI) were calculated as the fire's thermal 30 31 signature characterization. CI is a metric that integrates the temperature and the duration of 32 the burn (Pyle et al., 2015), which may characterize the heat effects on the soil better than 33 these parameters individually. For instance,  $CI_{100}$  integrates the area of the curve generated 34 over time in which the temperature exceeds 100°C during the heating phase and drops to 100°C during the cool-down phase. We calculated the CI by summing all the temperatures
between these two limits and multiplying the sum by the measurement time interval, as in Eq.
(1):

$$\mathrm{CI} = (\sum_{i=N}^{M} T_{i}) \cdot \Delta t$$

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Similarly, we calculated CI<sub>200</sub> and CI<sub>300</sub> according to the threshold temperatures of the soil
sample reaching 200°C and 300°C, respectively.

(1)

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#### 8 2.3. Laboratory methods

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10 Air-dried monoliths were divided into four layers: 0-1 cm, 1-2 cm, 2-3 cm, and 3-411 cm in depth. Each layer was weighed to calculate the bulk density. After hand sieving, the 12 fractions >2 mm and <2 mm were weighed separately to calculate the percentage of coarse 13 elements, so as to transform % (w/w) to t/ha.

14 Soil water repellency (WR) persistence and intensity were measured in the 2 mm hand-sieved samples. Soil WR persistence was determined by the water drop penetration time 15 (WDPT) method (Doerr, 1998). This test consists of placing drops of distilled water on the 16 soil surface and recording the time it takes for the water to penetrate completely into the soil. 17 The WR intensity was measured using the ethanol percentage (EP) test (Doerr et al., 2002; 18 19 Badía et al., 2012). Six drops of standardized solutions of ethanol in water (0.0, 5.0, 8.5, 13.0, 18.0, 24.0, and 36.0% v/v) were applied to the soil samples using a hypodermic syringe, and 20 21 their instant or short-term infiltration behavior was observed.

22 Soil pH was determined potentiometrically in a 1:2.5 ratio in H<sub>2</sub>O (McLean, 1982). The total carbon and total soil nitrogen contents were determined using an elemental 23 24 autoanalyzer (Vario Max CN Macro Elemental Analyser). Organic C (OC) was determined by the wet oxidation method (Nelson and Sommers, 1982). The inorganic Carbon (IC) was 25 26 measured volumetrically (with a calcimeter) after being treated with 6 M hydrochloric acid and expressed as equivalent CaCO<sub>3</sub> (Nelson, 1982). The electrical conductivity (EC) in a 1:5 27 ratio (soil:water) at 25°C was obtained as a measure of the total dissolved solutes (Rhoades, 28 1982). 29

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31 2.4 Statistical analysis

To identify the differences in the soil parameters due to the different treatments 1 2 (unburned, dry burned, and wet burned monoliths) and the sampling soil depth, a two-way factorial ANOVA (treatment x depth) was used. Assumptions of normality and 3 homoscedasticity were tested. Data that did not meet the prerequisites of the analysis were 4 transformed. In cases where, even after transformation, the prerequisites of normality were 5 not met, a Kruskal-Wallis test was used. In normally distributed data, the separation of means 6 7 was tested by Fisher's partial least-squares difference (PLSD) test with a significance level of p<0.01 or p<0.05. Differences between treatments at the same sampled depth were evaluated 8 9 by means of the t-test or Mann-Whitney U test. All these statistical analyses were conducted with StatView (SAS Institute Inc.). We also used a principal component analysis (PCA) of the 10 11 obtained data using XLSTAT software to investigate the relationships between the soil 12 properties over soil depth of the three treatments (unburned, dry burned, and wet burned). After standardization of the data, PCA was performed using the Pearson correlation matrix, 13 and a Varimax rotation was applied to the first two main components. 14

- 15
- 16 **3. Results**
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## 18 *3.1. Fire's thermal signature with soil depth*

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20 The highest temperatures (Tmax) were found at the soil surface and declined gradually with soil depth (Table 2). It can be observed a strong thermal gradient across a short thickness 21 (from the surface to a soil depth of 3 cm): from 590°C to 131°C in dry topsoil and from 22 324°C to 47°C in wet topsoil. The Tmax measured in dry soils was three to four times higher 23 than that in wet soils at each depth (Table 2). The effect of moisture content was also verified 24 in the duration of the heat at the different soil depths (Table 3). The maximum temperature 25 26 (Tmax) reached in the soil monoliths clearly depended on the initial soil water content (F=124.3; P<0.0001) and soil depth (F=51.3; P<0.0001). The interaction of soil humidity x 27 28 depth was also significant (F=15.6; P<0.0001).

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Table 2. Maximum temperatures (°C) reached in a mollic horizon burned in wet and dry conditions
(mean of three replicates for each moisture content and soil depth).

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Table 3. Heating duration (min) of range temperatures, maximum temperatures (Tmax in °C), and time (min) when Tmax was reached at four measurement points (surface and 1, 2, and 3 cm Ah soil

1 depth) of burned monoliths of a mollic horizon (mean of three replicates for each moisture content and2 soil depth).

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4 At the beginning of the heating process (minute 5), the temperature curves over time 5 have a lower slope in the burned wet soil than in the burned dry soil (Fig. 1); the water contained within the soil tends to hold back the soil heating until it is vaporized (Campbell et 6 7 al., 1995). At the end of heating, cooling is much faster in the wet-burned soil because the CIs 8 are lower for wet soils than for dry soils. The energy applied as heat is invested at the phase 9 change of the water, which is 2257 kJ/kg (539.4 cal/g) at 100°C (latent heat of vaporization of water), and it does not provide a thermal increase of the soil. In our experiment, the water 10 content was not exhausted in the wet soil after 15 minutes of burning, retaining approximately 11 300-400 g water/soil kg. However, the moisture content of air-dried soil (115-130 g water/ 12 soil kg) decreased to 30–42 g/kg after burning even at 3-4 cm soil depth (Appendix). 13

14 The heat in depth tended to last longer than on soil surface. For that reason, the surface 15 temperature curve overlaps the curves at depth (the first, second, and even third centimeter) 16 after removing the heat source (Fig. 1).

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Fig. 1. Temperature behavior on the surface and at the first centimeters of depth of a burned mollic topsoil under contrasted soil moisture content: air-dried versus field capacity. The flame was switched on at minute 5 and switched off at minute 20 (blue arrows). Although the temperature measurements were recorded for 210 min, in this figure, only the first 60 min are shown to remark on the main variations.

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24 The charred intenisties (CI<sub>100</sub>, CI<sub>200</sub>, and CI<sub>300</sub>) were clearly dependent on the soil water content and soil depth (Table 4). The interaction of soil humidity x soil depth was also 25 26 always significant, suggesting again the influence of soil water amount on heat transfer. The CI<sub>100</sub> values were positively and highly correlated (p<0.01) with the CI<sub>200</sub> and CI<sub>300</sub> values, as 27 well as with Tmax. Therefore, all CIs were negatively and significantly correlated with the 28 soil moisture content (Table 5). The strong decrease of these indexes with depth, which 29 30 characterizes the amount of heat input to the topsoil during burning, provides additional 31 evidence about the thermal inertia of the soil.

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**Table 4.** Charred intensity at different temperatures ( $CI_{100}$ ,  $CI_{200}$ , and  $CI_{300}$ ), in °C min, obtained in a mollic horizon burned in wet and dry conditions.

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Table 5. Correlation matrix (Pearson) between maximum temperature (Tmax) and charred intensities
(CI) for a mollic horizon burned in dry and wet conditions (n=24). Black values are significantly
different at 95% R>0.404 (at 99%, R>0.515).

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The maximum values of  $CI_{100}$  were found in the first centimeter of the dry soil (9,992) 6 7  $\pm$  1,286 °C min), whereas only 3,593  $\pm$  3,054 °C min was reached in the same soil in wet conditions. The CI values obtained by Pyle et al. (2015) in the pyrolysis of plant biomass for 8 charcoal production are, logically, many orders of magnitude  $(10^4)$  higher than the ones 9 obtained in this controlled burn, due to the differences in the durations of the process. 10 11 Variations in soil properties with heat can be linear from relatively low temperatures, e.g., mass yield, or show drastic decreases from a certain thermal threshold, e.g., C and N contents 12 (Pyle et al., 2015). In our work, the correlation of the three CIs with soil properties are similar 13 to the relation of the Tmax with these properties (Table 5), which indicates that this heat 14 duration don't give us additional information on soil effects by burning. 15

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17 3.2. Soil physical properties: soil moisture, water repellency (WR), and color

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The soil moisture content decreased significantly up to 4 cm depth after burning the 19 air-dried soil. Air-dried topsoil had 130 g of water/kg of soil, which decreased to 30 g of 20 water/kg of soil after burning through the first centimeter, and there was up to 40 g of 21 22 water/kg of soil below the first centimeter. Wet topsoil at field capacity contained approximately 510 g of water/kg of soil in the first centimeter, decreasing to 302 g of water/kg 23 24 of soil after burning. However the moisture content did not undergo major changes below. This is related to the temperatures above 100°C reached in dry soils, even at 3 cm depth. In 25 the wet soils, there were only temperatures higher than 100°C in the first centimeter of depth. 26

27 No significant differences in the soil moisture content were found in any sample depth after burning wet soils; although the moisture content in the first centimeter of wet-burned 28 29 soil (302 g/kg) was slightly lower than the content in the second centimeter (408 g/kg). We assume some loss of water by vaporization in this first centimeter, because the Tmax reached 30 for this centimeter ranged from 334°C to 124°C. However, apparently little soil water 31 vaporized below the first centimeter since the Tmax registered at a soil depth of 2 cm was 32 61°C. We have not detected any transport of evaporated soil moisture to cooler regions deeper 33 in the soil, referred to as "evaporative front" by Massman (2012). 34

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Burning significantly decreased the WR persistence (WDPT test) at the first, second, and third centimeter of the Ah horizon. The main changes were found in the first centimeter, where the high WDPT in air-dried unburned soil (923 seconds) disappeared after burning, thus turning the severely repellent topsoil into a hydrophilic topsoil; the severe repellency of the second and third centimeter decreased to strong repellency. Although dry soils maintain some WR level after burning in depth, WR is practically absent in the wet soils (after and before the fire).

Burning also significantly decreased the WR intensity (EP test) in the first, second and
third centimeter of the air-dried Ah horizon, converting the strong or very strong intensity of
the repellence into a very hydrophilic layer (first centimeter) or a strong intensity layer
(second and third centimeter). Only in the fourth centimeter did the EP remain unchanged, as
occurred with the WDPT. In fact, EP and WDPT data are positively and significantly
correlated (R=0.903; p<0.01) in this work.</li>

The color (Munsell code) was significantly darker in the first centimeter, but only after burning in dry conditions, from 10YR 5/2 to 10YR 3/1. No significant changes were detected at greater soil depths. The value or brightness decreased significantly from  $4.7\pm0.8$ (unburned) to  $2.7\pm0.6$  (dry burned); the chroma decreased not significantly from  $2.0\pm0.0$ (unburned) to  $1.3\pm0.6$  (dry burned). We obtained (Table 5) a highly significant and negative correlation (p<0.01) between the value (or brightness) and heating (both Tmax and CIs).

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# 22 3.3. Chemical properties: organic C, total N, EC, pH, and carbonates

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As expected, the OC content in the unburned topsoil samples progressively and 24 25 significantly (P=0.027) decreased with depth, from the first to the fourth centimeter. But 26 burning dry soil drastically reduced the highest OC content at the surface and eliminated the 27 usual OC decrease with soil depth. Specifically, the initial OC content in the first centimeter of the Ah horizon (143.55 g/kg) decreased significantly by approximately half (to 77.07 28 29 g/kg). In the burned wet soil, this decrease was also significant in the first centimeter of the mollic horizon, although it was less intense (to 104.57 g/kg). Burning lead to losses of 2.53 30 and 3.86 Mg TOC ha<sup>-1</sup> at 1-cm depth in the previously wet and dry soil monoliths, 31 respectively. This means that burning resulted in the loss of one-quarter of the pre-fire OC of 32 the wet monolith and one-half of the OC under the dry conditions of the upper centimeter of 33

the Ah horizon. Instead, soil burning only caused a (not significant) loss of one-seventh of the
pre-fire total N (TN) in the upper centimeter of the Ah horizon under wet conditions, but it
causes a significant loss of one-third of the total N under dry conditions. The original C/N
ratio (15.3 ± 2.5) decreased to 13.5 ± 1.0 for burning wet soils and 12.5 ± 0.9 for burning dry
soils in the first centimeter. Neither OC nor TN were significantly reduced at depths below 2
cm. The OC and TN contents are positively and significantly correlated (R=0.962; p<0.01);</li>
thus, much of the discussion regarding OC can be applied to TN.

8 The EC in the unburned mollic horizon had similar values across the upper 4 cms 9 (approximately 200 µS/cm). Burning resulted in an increase in EC, which was significant for 10 the uppermost centimeter when the soil was dry, but not when the soil was wet. Specifically, 11 EC in the first centimeter of unburned topsoil (220 µS/cm) was doubled for wet soil (433 12  $\mu$ S/cm) and increased by a factor of five for dry soil (1060  $\mu$ S/cm). Thus, the homogeneous EC distribution in the first 4 cm of topsoil (P=0.627) was replaced by a strong heterogeneity 13 due to the significant increase (P=0.004) in the concentration of soluble ions in the top of the 14 soil after burning in both situations, wet and dry soil conditions. 15

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17 Fig. 2. Effect of soil moisture on the OC content (in Mg/ha) and EC ( $\mu$ S/cm) of burned soils.

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The inorganic C, expressed as carbonate content, was very high (ranging from 35% to 47% from the first to the fourth centimeter of depth), which determines the basic reaction (pH ranging from 7.9 to 8.1) of the soil. Neither carbonates nor the pH were significantly affected by burning, unlike other properties (Table 6).

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Table 6. Significant effects of experimental burning of a mollic horizon under contrasted moisture
 content. The decrease is shown in red and the increase is shown in green. Properties without
 significant changes (pH, inorganic C, chroma) are not included.

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### 28 *3.4. Overall Analysis*

To investigate the relationships between the effects of controlled burning on wet and dry conditions and the soil properties affected at different soil depths (1, 2, 3, and 4 cm ), we used a PCA of the entire set of data and the replicates. The factor scores of the soil samples and the scores of the individual soil properties on the two first axes (which accounted for 62.6% of the variance) are displayed (Fig. 3). Based on the PCA results, the following can be seen:

- The soil samples (three replicates) are distributed according to the burning treatment 1 2 (unburned or burned), original moisture content (dry or wet), and soil depth (Fig. 3, above). Axis I had large positive loadings (right half) for the unburned soil samples. The distribution 3 of the soil samples along this axis was conspicuously related to depth. Thus, most samples at 4 the top (first and second centimeter) had large positive scores (on the right side of the graph) 5 corresponding to high total C, OC, TN, and WR, whereas the deeper soil samples had the 6 7 lowest scores (on the left of the graph), corresponding to higher inorganic C and less OC. Axis II had positive loadings (Fig. 3 above) for dry-burned soils, which are again distributed 8 9 along this axis in relation to their sampled depth: at the top, the most superficial soil samples and, at the bottom, the deepest soil samples. 10

The soil properties and heat signature (Tmax and CI) are shown in the lower part of Fig. 3.
The parameters related to heat, as well as the dissolved solutes (or EC), all of them close,
have very positive loadings (at the top of the graph). This finding mirrors the effect of the
controlled fire on air-dried soils, especially at the surface. Soil samples with negative loadings
(down and left) are not affected by fire, and analogously, the soil parameters are scarcely
affected by fire.

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Fig. 3. Biplot diagram of the first and second axes obtained from principal component analysis (PCA)
of the soil sample (above) and soil property (below) data. The first number indicates the replicate and
the second number is the soil depth sampled (cm).

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#### 22 **4. Discussion**

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# 24 *4.1. Fire's thermal signature with soil depth*

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The temperatures obtained in the experimental burn were within the range offered in a 26 recent review on prescribed burns (Cawson et al., 2012), with Tmax from 300°C to 900°C at 27 the litter layer, 100°C to 625°C at the litter-duff interface, 80°C to 210°C at the mineral soil 28 29 surface, and 25°C to 200°C at 2.5 cm soil mineral depth. However, in a previous review, Úbeda and Outeiro (2009) reported temperatures higher than 1,000°C on the ground surface 30 affected by wildfires; the temperatures seldom exceeded 100°C (e.g., under dense pine forest) 31 below a soil depth of 5 cm. DeBano et al. (1979) reported heating surface temperatures in 32 North American forests over 300°C and lasted for 5-10 min. In the shrublands and woodlands 33

of SE Australia, Bradstock and Auld (1995) found temperatures of 60°C-120°C in subsurface
soil (3 cm), with heating durations up to 20 min.

Fernández et al. (2013) investigated burned scrublands (mixed heathlands of Erica 3 australis L., ssp. aragonensis (Willk.) and Pterospartum tridentatum (L.) Willk.) in 4 experimental plots in NW Spain. They found extraordinarily low temperatures, with a mean 5 Tmax of 152°C (ranging from 70.0°C to 402.6°C) at the soil organic layer surface and a mean 6 7 Tmax at the mineral soil surface of 58.1°C (ranging from 11.5°C to 293.8°C); the soil was not heated at 2 cm below the mineral surface. Similarly, Vega et al. (2014), in an experimental 8 9 prescribed burning of shrublands dominated by gorse (Ulex europaeus L.) in NW Spain, found a mean Tmax of 162°C on the soil surface and a Tmax range of 22°C to 43°C at 2 cm 10 11 below the mineral soil surface. In these experimental burns, the heat did not noticeably 12 penetrate the soil because the fires only lasted for a short period and the soil moisture was high, approximately 60% w/w in the organic layers and 30%–40% w/w in a 0–5 cm mineral 13 horizon (Fernández et al., 2013; Vega et al., 2014). In addition, in a prescribed burn of the 14 understory of dry eucalyptus forests in the uplands of Victoria (Australia), Cawson et al. 15 (2016) found that the heating duration was very short, i.e., surface temperatures that were 16 greater than 400°C lasted for 2 seconds, and temperatures greater than 200°C lasted for 6 17 seconds. But in prescribed burns of thorny cushion dwarf Echinospartum horridum at the 18 Buisán location in the Pyrenees range (Girona et al., 2017), the residence time of heat was 19 20 higher than in understory burns; so, the temperature at the mineral soil surface reached 438°C and remained above 200°C for 6.5 min; however, there were negligible increases at a soil 21 22 depth of 1 cm. In contrast, at the neighboring Tella location (Armas-Herrera et al., 2016), and with a similar prescribed burn, the temperature at 1 cm soil depth reached nearly 400°C and 23 24 remained above 200°C for 25 min. This resulted in a heat signature with a strong impact, especially on the soil biological properties. Cawson et al. (2016) carried out prescribed burns 25 26 of the shrubby or grassy understory of dry Eucalyptus forests at different locations in the 27 uplands of Victoria; these had low moisture values, both in the surface litter (10%-15%) and 28 for 0–5 cm of mineral topsoil (7%–9%). They found a Tmax of 969°C at the surface and peak temperatures of 728°C at shallow soil depths (0.5 cm), but they seldom detected changes in 29 30 temperature at a depth of 4 cm. Below the surface, the peak temperatures were lower but the 31 heating durations were longer, with temperatures over 100°C lasting 23 min on average. The 32 same authors remarked on the spatial heterogeneity of temperatures in prescribed burned 33 landscapes. Also, Santín et al. (2016), in a high-intensity experimental fire of a Canadian 34 boreal forest of jack pine (Pinus banksiana), found that sensors placed only 2 m apart

registered Tmax values ranging from 550°C to 976°C at the forest floor surface. Even in
experimental plots using the backfire burning technique (against the wind to favor slow fire
propagation and fuel combustion), litter and duff consumption was irregular (Vega et al.,
2014).

5 Dry soil acts as a thermal insulator, resulting in slow rates of heat transfer within the soil; 6 however, by the same principle, these insulating properties make it more difficult for the soil 7 to lose heat and it is retained longer (Massman, 2012). The slow heating and rapid cooling 8 observed in burned wet soils are related to the heat capacity and the heat conductivity, which 9 increase in soils at field capacity, where water fills most of the soil pores (DeBano et al., 10 1998).

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#### 12 4.2. Soil physical properties: soil moisture, water repellency (WR) and color

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WR (both WDPT and EP) naturally occurs in unburned air-dried soils. This naturally very high "background" level of repellency was previously reported for mollic horizons (Badía et al., 2014b) and, to a lesser extent, for ochric horizons (Aznar et al., 2016) under an Aleppo pine forest (*Pinus halepensis* L.), as well as for other soil types under different vegetation (Doerr et al., 2006; Jordán et al., 2010; Keesstra et al., 2017). WR decreased progressively with soil depth (as reported by Badía et al., 2013a; Bodí et al., 2013; Benito et al., 2016; Keizer et al., 2008; among others).

Although the mollic horizon is extremely water repellent when dry, it is wettable at 21 field capacity, at practically any measured thickness (0 to 4 cm of soil depth). The burning of 22 the wet soil can not dry the soil, which stays wettable (Appendix). Soils exceeding a certain 23 24 threshold of moisture content may shift from water-repellent to wettable (DeBano, 2000). Bodí et al. (2013) found a critical soil moisture threshold above which the soil becomes 25 26 wettable, ranging from 5% (sandy soils) to 30% (clay soils). Benito et al. (2016) found that 27 the moisture range defining the presence or absence of repellency under field conditions was 28 22%–57% in pine plantations in NW Spain. It is common that the soil surface will be largely wettable in wet seasons and extremely water repellent during dry seasons (Benito et al., 2016; 29 30 Bodí et al., 2013; Keizer et al., 2008; Keesstra et al., 2017).

A fire-induced decrease in WR has been found in depth, up to 2–3 cm, where a Tmax range of  $241^{\circ}C-132^{\circ}C$  (and a  $CI_{100}$  range of  $4,618-2,130^{\circ}C$  min) was measured. According to DeBano et al (1998) this temperature range should have formed a water-repellent layer either in place or in underlying soil layers. The conspicuous elimination of WR without a

significant decrease in total organic C content could be related to the fragmentation by heat of 1 the long-chain molecules of the alkylic series: n-alkanes, alkenes, fatty acids and methylated 2 fatty acids (Badía et al., 2014b). Fire-induced reduction in WR is usually related with high 3 temperatures (Jordán et al., 2010; Giovannini, 2012; Keesstra et al., 2017). For instance, 4 Giovannini (2012) found that temperatures above 550°C completely destroyed soil 5 hydrophobicity. Jordán et al. (2010) showed that soil water repellency decreased after intense 6 7 burning of a Mediterranean heathland (SW Spain); these decreases were strong at the soil surface but diminished progressively with soil depth, as was observed in our mollic horizon. 8 9 Keesstra et al. (2017) found a high WR on the surface of long-unburned soils (>50 years unburned) in a Pinus halepensis forest from Mount Carmel range; in contrast, burned and 10 11 twice-burned areas have low and negligible soil WR, respectively, demonstrating that 12 recurrent fires reduce topsoil WR.

However, fire-induced reduction in WR is not always observed and, conversely, many 13 other studies have reported increases in WR after heating. For example, heating soil samples 14 at 205°C for 10 min or at 305°C for 5 min increased water repellency (DeBano et al., 1998); 15 the same authors indicates that the exposure to 480°C completely eliminated hydrophobicity. 16 Robichaud and Hungerford (2000) described millimetric displacements of the WR layer and 17 reported a WR increase at a soil depth of 2–5 cm, although the temperatures remained below 18 260°C. Vadilonga et al. (2008) found that WDPT increased moderately at the soil surface in 19 20 the plots exposed to high burn severity, and decreased slightly when the burn severity was low. Atanassova and Doerr (2011) observed an increase in soil WR after heating (at 300°C 21 22 for 10 min) eucalyptus forest topsoils (0–5 cm depth) with a sandy texture. Cawson et al. (2016) found that slight soil heating (temperatures about 200°C at the surface, lasting an 23 24 average of 6 seconds), during prescribed burns of shrubby or grassy understories of dry *Eucalyptus* forests, was related to the strength of soil water repellency at the soil surface. The 25 26 effect of heating on soil WR definitely depends on the fire signature and soil moisture content, among other soil properties and components (Cawson et al, 2012; Bodí et al., 2013; 27 28 Jordán et al., 2010; Benito et al., 2016; Prat-Guitart et al., 2016).

The darkening of the first centimeter by burning in dry conditions is related to the charring of soil organic matter (Ulery and Graham, 1993). Heating soils for 30 min in a muffle furnace at 25°C, 150°C, 250°C, and 500°C linearly decreased significantly (p<0.01) the luminosity or brigthness (R=-0.58) and even the chroma (R=-0.75) in an ochric calcareous horizon (Badía and Martí, 2003a). Instead, Cancelo et al. (2014), heating sieved samples of umbric horizons, found an increase in the luminosity and chroma of burned soils, which is related to a decrease in the amount of organic matter. This relation between organic matter and color in burned soils could be a useful tool to evaluate the evolution of burned landscapes by VIS-NIR-SWIR spectroscopy (Rosero-Vlasova et al., 2016). In other cases, reddening and yellowing of soils after burning have been related to the transformation of iron oxides at high temperatures as well as OC losses. For instance Ulery et al (2017) found strong changes in soil color (hue, value, and chroma) at soil depths up to 15 cm, after severe fires under burned slash piles or fallen logs.

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#### 4.3. Chemical properties: organic C, total N, EC, pH, and carbonates

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11 In our previous studies, in which a mollic horizon was burned in a similar way but with preservation of the upper O horizon (2.5 cm thick), the decrease in OC was not 12 significant in the Ah horizon, although the O horizon was partially transformed by 13 combustion into ashes and charred litter (Badía et al., 2014b). Its burning resulted in the loss 14 of two-thirds of the pre-fire OC of the O horizon and one-third of the OC of the upper 15 centimeter of the Ah horizon (Badía et al., 2014b). In absolute terms, O burning implies much 16 higher C losses, with a total loss of 13.1 Mg of TOC ha<sup>-1</sup> (Badía et al., 2014b). Our results, 17 obtained in controlled laboratory experiments, were quite similar to reported losses by Certini 18 et al (2011) in the field in sandy soils from pine forests affected by severe wildfires (14.3-19 17.8 Mg TOC ha<sup>-1</sup>). The C losses were attributed to the complete removal of the O horizon, 20 21 whereas the underlying Ah horizon soil did not show any significant change in its carbon 22 content.

With similar experimental burning in an ochric horizon with a thinner O horizon (1 cm thick), the decrease in OC in the first centimeter of the Ah was significant (Aznar et al., 2016). In this case, burning resulted in the loss of three-quarters of the pre-fire OC of the O horizon and one-half of the OC of the upper centimeter of the Ah horizon, which amounted to 3.6 Mg TOC ha<sup>-1</sup> (Aznar et al., 2016). We have not found significant soil OC losses in the second burned centimeter of the Ah horizon, neither in the mollic horizon nor in the ochric, with or without an overlaying O horizon.

The OC losses depend mostly on the duration of heating and the temperatures reached during burning (DeBano et al., 1979). As summarized by Certini (2005), there is partial consumption of organic matter at 200°C–250°C and complete consumption at 460°C. In laboratory experiments, we found C losses of approximately 90% in different Ah horizons heated 30 min to 500°C, while heating to 250°C caused a reduction of approximately 15%–

20% (Badía and Martí, 2003a,b); the progressive increase in temperatures caused a decrease 1 in the OC content and many other related properties, such as cation exchange capacity (CEC), 2 soil aggregation stability, microbial biomass, and soil respiration (Badía and Martí, 2003a,b). 3 In this sense, Ulery et al. (2017), studying the effect of severe, high-intensity burning under 4 heavy fuel loads in five wooded sites of California, found an extreme loss of OC and 5 therefore a large decrease of CEC, even at a soil depth of 15 cm. Hinojosa et al. (2016), in an 6 7 experimental burning of Cistus-Erica shrubland in central Spain, found that soil organic matter content was significantly lower in the burned soils  $(7.43 \pm 0.61\%)$  than in the unburned 8 9 ones  $(9.75 \pm 0.55\%)$ , collecting the top 5 cm of the soil. In this case, the average Tmax was 710°C at the soil surface, and the mean residence time was 13.5 min above 100°C. 10

Armas-Herrera et al. (2016), in a prescribed burning of thorny shrubs (Tella, NE 11 12 Spain), found a decrease in total soil OC content up to 3 cm (and total N up to 1 cm). This was due to an exceptional thermal signature because, as we mentioned above, the temperature 13 at a soil depth of 1 cm reached nearly 400°C and remained above 200°C for 25 min. San 14 Emeterio et al. (2016) studied soil N changes in a prescribed burning of Ulex gallii shrubland 15 in the western Pyrenees (NE Spain), sampling 10 cm of soil thickness. They found that the 16 total N content decreased significantly just after burning and it stayed lower in burned soils 17 than unburned soils two years later. This TN decrease was followed by an increase in mineral 18 N content. In this experimental burning, the highest temperature recorded was 71°C at a soil 19 20 depth of 2 cm. Instead, the organic matter and total N usually increase in burned soils if external black ashes and vegetation inputs are sampled with the mineral topsoil (Jiménez-21 22 González et al., 2016; Yusiharni and Gilkes, 2012). These ashes and charcoal, with mineral topsoil, can be quickly lost after erosive rainstorms following the fire due to transport with the 23 24 runoff (Badía and Martí, 2008; Hosseini et al., 2016).

A sharp increase in the pH of soil samples heated to 500°C or higher has been 25 26 commonly observed in various mineral soils (Badía and Martí, 2003a; Giovannini, 2012) probably related to thermal decomposition of calcite to CaO (Rodríguez-Navarro, et al., 2009). 27 28 But in this study, neither carbonates nor pH were significantly affected by burning because, temperatures lower than 500°C were recorded at a soil depth of 1 cm: 439°C in dry monoliths 29 30 and 124°C in wet monoliths. Only dry monoliths, in their most superficial millimeters, can 31 exceed 500°C for some seconds, which is not enough to modify the pH or the carbonates 32 content.

33 Some prior works have reported an increase in topsoil pH (as well as EC) in high-34 intensity fires, because the combustion of the plant biomass and/or organic layer released ash and charcoal that were rich in base cations (Certini, 2005; Pereira et al., 2012; Martín et al.,
2013; Heyradi et al., 2016; Ulery et al., 2017, among others). However, Marcos et al. (2009),
in a low-intensity prescribed fire of heathlands in three acid soils (pH<4.3) in the Cantabrian</li>
mountains (NW Spain), observed no significant differences between the pH in burned and
control soils (0–5 cm soil thickness), in spite of the extreme basicity (pH 8.5–9.5) of the ashes
covering the mineral soil.

The increase in EC or total dissolved solutes, in burned dry soil is not related to release of ash from burned plant biomass, so it should be related to the mineralization of the soil organic matter by heating. In our previous studies, the EC increased significantly heating ochric calcareous samples for 30 min at temperatures above 250°C (Badía and Martí, 2003a). The original EC of the saturated extract paste (ECe) of this ochric horizon (0.9 dS/m) increased slightly to 1.2 dS/m at 150°C and significantly to 3.4 dS/m at 250°C, decreasing to 2.4 dS/m by heating at 500°C when calcium precipitates are formed (Badía and Martí, 2003a).

Definitely, soil thickness directy affected by burning is mainly limited to the first cm, either wet or dry. The increase in the heat capacity of the soil, due to its high water content, "protect" some physical and chemical properties. However, this high water content can increase the heat conductivity of the soil and transfer soil-sterilizing temperatures by the porous phase to a greater soil depth. This would affect some biological soil properties, which are more sensitive than the physical and chemical properties (Santín and Doerr, 2016; Hinojosa et al., 2016; García-Orenes et al., 2017).

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### 22 5. Conclusions

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Tmax and the heating duration are significantly lower in the wet soil than in air-dried soil in the first and second centimeters of depth. At field capacity, water fills a significant volume fraction of the soil medium, which increases the heat capacity of the soil. This increase allows the soil to absorb certain amount of heat without a subsequent temperature increase, which slows the heat transmission to deeper soil layers. Once the fire is extinguished, the cooling is faster in the still wet soil than in dry soil due to the higher heat conductivity of the former.

The original high WR in unburned air-dried soil, both persistence and intensity, decreased significantly to the third centimeter of soil depth by burning. OC and total N losses were higher by burning soil in air-dried conditions than in wet conditions, but only in the upper topsoil centimeter. Neither OC nor TN were significantly reduced by burning at soil depths below 2 cm. The remaining organic matter was intensely charred, which blackens the first
 centimeter of air-dried soils, reducing its brightness or luminosity.

As a result of OC oxidation, the soluble ions (EC) significantly increased due to burning in 3 the first centimeter of air-dried soils. Both the pH and carbonates do not vary significantly, 4 even in the first centimeter. The effects of burning soil at field capacity were shallower and 5 less intense than in the air-dried conditions, proving that soil moisture provides a certain 6 7 degree of protection against fire; this should be taken into account when performing a prescribed burn. In any case, we provide evidence of changes just for the most superficial 8 9 centimeter of soil depth in some soil physical and chemical properties, which are less sensitive to heat than biological properties. Complementary research is needed in order to 10 11 identify the soil moisture effects in other soil properties and soil types after burning.

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21

## 22 **References**

23

- Aznar, J.M., González-Pérez, J.A., Martí C., Badía-Villas D., 2016. At what depth are the
   properties of gypseous forest soil affected by fire? Land Degrad. Dev. 27, 1344–1353.
- Badía, D., Martí, C., 2003a. Plant ash and heat intensity effects on chemical and physical
  properties of two contrasting soils. Arid Land Res. Manag. 17, 23-41.
- Badía, D., Martí, C., 2003b. Effect of simulated fire on organic matter and selected
   microbiological properties of two contrasting soils. Arid Land Res Manag. 17, 55-69.
- Badía, D., Martí, C., 2008. Fire and rainfall energy effects on soil erosion and runoff
  generation in semi-arid forested lands. Arid Land Res. Manag. 22, 93-108.

Armas-Herrera, C.M., Martí, C., Badía D, Ortiz-Perpiñá, O., Girona-García, A., Porta, J.,
 2016. Immediate effects of prescribed burning in the Central Pyrenees on the amount and
 stability of topsoil organic matter. Catena 147, 238–244.

- Badía, D., Aguirre, A.J., Martí, C., Márquez, M.A., 2013a. Sieving effect on the intensity and
   persistence of soil water repellency: A case study using different soil depths and soil
   types from NE-Spain. Catena 108, 44-49.
- Badía, D., Martí C., Aznar, J.M., León F.J., 2013b. Influence of slope and parent rock on soil
  genesis and classification in semiarid mountainous environments. Geoderma 193-194,
  13-21.
- Badía, D., Martí, C., Aguirre, A.J., Aznar, J.M., González-Pérez, J.A., De la Rosa, J.M.,
  León, F.J., Ibarra P., Echeverría M.T., 2014a. Wildfire effects on nutrients and organic
  carbon of a Rendzic Phaeozem in NE Spain: changes at cm-scale topsoil. Catena 113,
  267-275.
- Badía, D., González-Pérez, J.A., Aznar, J.M., Arjona-Gracia, B., Martí-Dalmau, C., 2014b.
  Changes in water repellency, aggregation and organic matter of a mollic horizon burned
  in laboratory: soil depth affected by fire. Geoderma 213, 400-407.
- Benito, E., Rodríguez-Alleres, M., Varela, M.E., 2016. Environmental factors governing soil
  water repellency dynamics. Land Degrad. Dev. 27, 719–728.
- Bento-Gonçalves, A., Vieira, A., Úbeda, X., Martin, D., 2012. Fire and soils: Key concepts
  and recent advances. Geoderma 191, 3-13.
- Bodí, M.B., Muñoz-Santa, I., Armero, C., Doerr, S.H., Mataix-Solera, J., Cerdà A., 2013.
  Spatial and temporal variations of water repellency and probability of its occurrence in calcareous Mediterranean rangeland soils affected by fires. Catena 108, 14–25.
- Bradstock RA, Auld TD. 1995. Soil temperature during experimental bushfire in relation to
  fire intensity: consequences for legume germination and fire management in southeastern Australia. J. Appl. Ecology, 32: 76–84.
- Campbell, G.S., Jungbauer, J.D., Bristow, K.L., Hungerford R.D., 1995. Soil temperature and
  water content beneath a surface fire. Soil Sci. 159 (6), 363–374.
- Cancelo J., Cachaldora C., Díaz-Fierros F., Prieto B., 2014. Colourimetric variations in burnt
   granitic forest soils in relation to fire severity. Ecol. Indic. 46, 92-100.
- Cawson, J. G., Sheridan, G. J. Smith H. G. and Lane, P. N. J., 2012. Surface runoff and
  erosion after prescribed burning and the effect of different fire regimes in forests and
  shrublands: a review. Int. J. Wildland Fire, 21, 857–872.
- Cawson, J.G., Nyman, P., Smith, H.G., Lane, P.N.J., Sheridan, G.J., 2016. How soil
   temperatures during prescribed burning affect soil water repellency, infiltration and
   erosion. Geoderma 278, 12–22.
- Certini G., 2005. Effects of fire on properties of forest soils: a review. Oecologia 143, 1–10.

1	Certini, G., Nocentini C., Knicker, H. Arfaioli P. Rumpel, C., 2011, Wildfire effects on soil
2	organic matter quantity and quality in two fire-prone Mediterranean pine forests
3	Geoderma 167-168, 148–155.
4	De Partearroyo, R., López, M., Viorreta, J.L., Cuevas J., 2012. Resumen de actuaciones del
5	EPRIF de Aragón, in González Pan, J.R. (Ed.), Los EPRIF. Planificación y desarrollo de
6	las actuaciones. Ministerio de Agricultura, Alimentación y Medio Ambiente, Madrid, pp.
7	330–355.
8	DeBano, L.F., 2000. The role of fire and soil heating on water repellency in wildland
9	environments: a review. J. Hydrol. 231-232, 195-206.
10	DeBano, L.F., Rice, R.M., Conrad, E.C., 1979. Soil heating in chaparral fires: effects on soil
11	properties, plant nutrients, erosion, and runoff. USDA Forest Service, Pacific Southwest
12	Forest and Range Experiment Station, Research Paper PSW-145 (Berkeley, CA).
13	DeBano, L.F., Neary, D.G., Ffolliott, P.F., 1998. Fire Effects on Ecosystems. John Wiley &
14	sons. USA. 333 pp.
15	Doerr, S.H., 1998. On standardizing the 'water drop penetration time' and the 'molarity of an
16	ethanol droplet' techniques to classify soil hydrophobicity: a case study using medium
17	textured soils. Earth Surf. Process. Landf. 23, 663-668.
18	Doerr, S.H., Dekker, L.W., Shakesby, R.A., Ritsema, C.J., Bryant R., 2002. Water repellency
19	of soils: the influence of ambient relative humidity. Soil Sci. Soc. Am. J. 66, 401-405.
20	Doerr, S.H., Shakesby, R.A., Blake, W.H., Chafer, C.J., Humphreys G.S., Wallbrink P.J.,
21	2006. Effects of differing wildfire severities on soil wettability and implications for
22	hydrological response. J. Hydrol. 319, 295-311.
23	Fernández, C., Vega, J.A., Fonturbel, T., Jiménez, E., Pérez, J.R., 2008. Immediate effects of
24	prescribed burning, mastication and clearing on runoff, infiltration and erosion in a
25	shrubland area in Galicia (NW Spain). Land Degrad. Dev 19, 502–515.
26	Fernández, C., Vega, J.A., Fonturbel, T., 2013. Fuel reduction at a Spanish heathland by
27	prescribed fire and mechanical shredding: Effects on seedling emergence. J. Environ.
28	Manage. 129, 621-627.
29	García-Orenes, F., Arcenegui, V., Chrenkova, K., Mataix-Solera, J., Moltó, J. Jara-Navarro
30	A.B., Torres, M.P. 2017. Effects of salvage logging on soil properties and vegetation
31	recovery in a fire-affected Mediterranean forest: A two year monitoring research. Science
32	of the Total Environment 586, 1057–1065.
33	Giovannini, G., 2012. Fire in agricultural and forestal ecosystems on soil. Edizioni ETS. 86
34	pp. Pisa, Italy.

Girona-García, A., Badía, D., Martí, C., Ortiz-Perpiñá O., Mora, J.L., Armas-Herrera C.M.,
 2017. Effects of prescribed fire for pasture management on soil organic matter and
 biological properties: a 1-year study case in Buisán, Central Pyrenees. Land Degrad. Dev
 (submitted).

- Keizer, J.J., Doerr, S.H., Malvar, M.C., Prats, S.A., Ferreira, R.S.V., Oñate, M.G., Coelho,
  C.O.A., Ferreira, A.J.D., 2008. Temporal variation in topsoil water repellency in two
  recently burnt eucalypt stands in north-central Portugal. Catena 74, 192–204.
- Jordán, A., González, F.A., Zavala, L.M., 2010. Re-establishment of soil water repellency
  after destruction by intense burning in a Mediterranean heathland (SW Spain). Hydrol.
  Process. 24, 736–748.

Hinojosa M.B., Parra A, Laudicina V.A., Moreno, J.M., 2016. Post-fire soil functionality and
 microbial community structure in a Mediterranean shrubland subjected to experimental
 drought. Sci. Total Environ. 573, 1178–1189

Hosseini, M., Keizer, J.J. Gonzalez-Pelayo, O., Prats, S., Ritsema, C., Geisse, V., 2016. Effect
of fire frequency on runoff, soil erosion, and loss of organic matter at the micro-plot scale
in north-central Portugal. Geoderma 269, 126-137.

Jiménez-González, M.A., De la Rosa, J.M., Jiménez-Morillo N.T., Almendros, G., González Pérez, J.A., Knicker, H., 2016. Post-fire recovery of soil organic matter in a Cambisol from
 typical Mediterranean forest in Southwestern Spain. Sci. Total Environ. 572, 1414-1421.

Keesstra, S., Wittenberg, L., Maroulis, J., Sambalino, F., Malkinson, D., Cerdà, A., Pereira,
 P., 2017. The influence of fire history, plant species and post-fire management on soil

water repellency in a Mediterranean catchment: The Mount Carmel range, Israel. Catena
149, 857–866.

- Marcos, E., Villalón, C., Calvo L., Luis-Calabuig, E., 2009. Short-term effects of
  experimental burning on soil nutrients in the Cantabrian heathlands. Ecol. Eng. 35, 820–
  828.
- Massman, W.J., 2012. Modeling soil heating and moisture transport under extreme
  conditions: Forest fires and slash pileburns, Water Resour. Res., 48, W10548,
  doi:10.1029/2011WR011710.
- 30 McLean, E.O., 1982. Soil pH and lime requirement. In Methods of Soil Analysis. Part 2:
- Chemical and Microbiological Properties, Page AL, Miller, RH, Keeney DR (eds). 2nd
- edition. American Society of Agronomy. Madison, Wisconsin; 199–224

21

1	Nelson, R.E., 1982. Carbonate and gypsum, in: Page AL, Miller RH, and Keeney DR (Eds.),
2	Methods of Soil Analysis. Part 2: Chemical and Microbiological Properties, second ed.
3	American Society of Agronomy. Madison, Wisconsin, pp. 181-198.
4	Nelson, R.E., Sommers L.E., 1982. Total carbon and organic matter, in: Page AL, Miller RH,
5	and Keeney DR (Eds.), Methods of Soil Analysis. Part 2: Chemical and Microbiological
6	Properties, second ed. American Society of Agronomy. Madison, Wisconsin, pp. 539-557.
7	Pereira, P., Úbeda, X., Martin, D.A., 2012. Fire severity on ash chemical composition and
8	water-extractable elements. Geoderma 191, 105-114.
9	Prat-Guitart N., Rein G., Hadden R.H., Belcher C.M., Yearsley J.M., 2016. Effects of spatial
10	heterogeneity in moisture content on the horizontal spread of peat fires. Sci. Total Environ.
11	572, 1422-1430.
12	Pyle, L.A., Hockaday, W.C., Boutton, T., Zygourakis, K., Kinney, T.J., Masiello C.A., 2015.
13	Chemical and isotopic thresholds in charring: Implications for the interpretation of
14	charcoal mass and isotopic data. Environ. Sci. Technol. 49 (24), 14057–14064.
15	Rhoades, J.D., 1982. Soluble salts, in: Page AL, Miller RH and Keeney DR (Eds.), Methods
16	of Soil Analysis. Part 2: Chemical and Microbiological Properties, second ed. American
17	Society of Agronomy. Madison, Wisconsin, USA, pp. 167-180.
18	Robichaud, P.R Hungerford, R.D., 2000. Water repellency by laboratory burning of four
19	northern Rocky Mountain forest soils. J Hydrol. 231–232, 207-219.
20	Rodríguez-Navarro, C., Ruiz-Agudo, E., Luque, A., Rodríguez-Navarro, A.B., Ortega-Huerta
21	M., 2009. Thermal decomposition of calcite. American Mineralogist 94, 578-593.
22	Rosero-Vlasova, O., Pérez-Cabello, F., Montorio, R., Vlassova, L., 2016. Assessment of
23	laboratory VIS-NIR-SWIR setups with different spectroscopy accessories for
24	characterization of soils from wildfire burns. Biosystems Engin. 152: 51-67.
25	San Emeterio, L., Múgica, L., Ugarte, M.D., Goicoa, T., Canals, R.M., 2016. Sustainability of
26	traditional pastoral fires in highlands under global change: Effects on soil function and
27	nutrient cycling. Agric. Ecosyst. Environ. 235, 155-163.
28	Santín, C., Doerr, S.H., 2016 Fire effects on soils: the human dimension. Phil. Trans. R. Soc.
29	B 371, 20150171. http://dx.doi.org/10.1098/rstb.2015.0171
30	
31	Santín, C., Doerr, S.H., Merino, A., Bryant R., Loader N.J., 2016. Forest floor chemical
	Santín, C., Doerr, S.H., Merino, A., Bryant R., Loader N.J., 2016. Forest floor chemical transformations in a boreal forest fire and their correlations with temperature and heating
32	Santín, C., Doerr, S.H., Merino, A., Bryant R., Loader N.J., 2016. Forest floor chemical transformations in a boreal forest fire and their correlations with temperature and heating duration. Geoderma 264, 71–80.
32 33	<ul> <li>Santín, C., Doerr, S.H., Merino, A., Bryant R., Loader N.J., 2016. Forest floor chemical transformations in a boreal forest fire and their correlations with temperature and heating duration. Geoderma 264, 71–80.</li> <li>Úbeda, X., Outeiro L., 2009. Physical and chemical effects of fire on soil. In: A. Cerdà and</li> </ul>

- 1 Inc. Eds. Enfield, New Hampshire, USA, pp. 105-132.
- 2 Ulery, A.L., Graham, R.C., 1993. Forest fire effects on soil color and texture. Soil Sci. Soc.
  3 Am. J. 57, 135-140.
- Ulery, A.L., Graham R.C., Gotofth, B.R., Hubbert K.R., 2017. Fire effects on cation
  exchange capacity of California forest and woodland soils. Geoderma 286, 125–130.
- Vadilonga, T., Ubeda X., Germann, P.F., Lorca M., 2008. Effects of prescribed burnings on
  soil hydrological parameters. Hydrol. Process. 22, 4249–4256.
- 8 Vega, J.A., Fernández, C., Fonturbel, T., González-Prieto S., Jiménez, E., 2014. Testing the
- 9 effects of straw mulching and herb seeding on soil erosion after fire in a gorse shrubland.
- 10 Geoderma 223–225, 79–87.
- 11 Yusiharni, E., Gilkes, R.J. 2012. Changes in the mineralogy and chemistry of a lateritic soil
- due to a bushfire at Wundowie, Darling Range, Western Australia. Geoderma, 191: 140-
- 13 150.

Soil properties	Mollic horizon
Colour (dry)	10YR 5/1
Colour (moist)	10YR 3/1
Soil aggregate stability (%, w/w)	95
Bulk density $/(kg/m^3)$	790
Field capacity (%, w/w)	36.2
Textural classes (USDA)	Clay Loam (22, 44, 34) <sup>a</sup>
Gravels (% w/w)	18.7
pH (H <sub>2</sub> O) 1:2.5	8.0
CaCO <sub>3</sub> equivalent (%, w/w)	49
ECe (dS/m)	1.8
Soil organic matter (%, w/w)	9.8
C/N ratio	15.9
Cation exchange capacity (cmol <sub>c</sub> /kg)	27.7

Main physical and chemical properties of the experimental mollic Ah horizon (0-30 cm)

Abbreviations: ECe, Electrical conductivity of saturated paste extract, <sup>a</sup>The percentage of each main fraction is in parentheses: sand, 2-0.05 mm; silt, 0.05-0.002 mm; clay, <0.0002 mm.

Maximum temperatures (°C) reached in a mollic horizon burned in wet and dry conditions (mean of three replicates by each moisture content and soil depth).

	Soil depth									
Burning in	Surface	1 cm	2 cm	3 cm	Р					
Dry	589.0a	426.8b	234.7c	130.9c	<0.001					
Wet	324.4a	123.3b	60.3c	47.3c	<0.044					
p=	0.082	0.006	0.023	0.057						

Lowercase letters compare between soil depths (within each line); *p* in the last row compares between dry and wet soil, within each soil depth (column)

Heating duration (min) of range temperatures, maximum temperatures (Tmax in °C), and time (minute) when Tmax has been reached at four measurement points (surface, and 1, 2 and 3 cm Ah soil depth) of burned monoliths of a mollic horizon (mean of three replicates by each moisture content and soil depth).

		Dry b	urned		Wet b	urned		
Soil depth (cm)	Surface	1	2	3	Surface	1	2	3
T max (°C)	589.0	426.8	234.7	130.9	324.4	123.3	60.3	47.3
Time (min)	9.7	15.3	19.9	20.6	17.8	20.2	20.8	24.5
Duration (min)								
0-100 °C	179.8	180.3	181.0	189.1	192.8	202.2	208.4	208.4
100-200	10.2	9.5	19.2	19.3	2.6	6.2		
200-300	3.5	6.5	8.2		12.3			
300-400	1.5	7.3			0.7			
400-500	3.2	4.8						
500-600	10.2							

Charred Intensities at different temperatures (CI 100, CI 200 and CI 300), in °C min, obtained in a mollic horizon burned in wet and dry conditions

			Soil depth							
CI 100 (°C min)	Burning in		Surface	1 cm	2 cm	3 cm				
	Dry		9992ab	7599b	4618bc	2130c				
	Wet		3593a	722b	0	0				
		<i>p</i> =	0.1552	0.0013	-	-				
CI 200 (°C min)	Burning in		Surface	1 cm	2 cm	3 cm				
	Dry		8676a	6316a	2101b	0				
	Wet		3177a	228b	0	0				
		<i>p</i> =	0.0396	<0.0001						
CI 300 (°C min)	Burning in		Surface	1 cm	2 cm	3 cm				
· · · · ·	Dry		7691a	4740b	563c	0				
	Wet		2388	0	0	0				
		p=	0.1648	-						

Lowercase letters compare between soil depths (within each line); *p* in the last row compares between dry and wet soil, within each soil depth (column)

Correlation matrix (Pearson) between Maximum Temperature (Tmax) and Charred Intensities (CI) for mollic horizon burned in dry and wet conditions (n=24). Black values are significantly different at P=95% R>0,404 (R>0,515 for P=99%).

-	Tmax	Ch	Soil		
Properties		<b>CI-100</b>	CI-200	CI-300	Moisture
T max (°C)	1				
CI-100 (°C min)	0,955	1			
CI-200 (°C min)	0,964	0,952	1		
CI-300 (°C min)	0,942	0,899	0,968	1	
Soil Moisture (%)	-0,611	-0,734	-0,579	-0,519	1
Organic C (%)	0,213	0,063	0,163	0,156	0,471
Total N (%)	0,266	0,124	0,204	0,185	0,401
Total C (%)	0,341	0,195	0,280	0,261	0,360
C/N ratio	-0,189	-0,238	-0,148	-0,111	0,319
Inorganic C (%)	-0,336	-0,254	-0,291	-0,243	-0,275
pH (1:2.5)	-0,413	-0,426	-0,371	-0,242	0,422
EC 1:5 (µS/cm)	0,812	0,808	0,849	0,866	-0,436
WDPT (class)	0,085	0,249	0,063	-0,046	-0,636
Ethanol (class)	-0,011	0,118	-0,049	-0,118	-0,635
Value (dry)	-0,767	-0,722	-0,754	-0,802	0,328
Chroma (dry)	-0,383	-0,437	-0,352	-0,330	0,511

Soil depth	Soil moisture content		Wa repel	nter lency	Colour	r: value	Orga	nic C	Tot	al N	Total d	issolved utes
(cm)	Dry	Wet	Dry	Wet	Dry	Wet	Dry	Wet	Dry	Wet	Dry	Wet
0-1				zero								
1-2				zero				1 1 1 1		1 1 1 1		
2-3				zero						1 1 1 1		
3-4				zero								

Table 6. Significant effects of experimental burning of a mollic horizon beneath contrasted moisture content: the decrease, in red, and the increase, in green. Properties without significant changes are not indicated (pH, inorganic C, chroma)





Fig. 2. Effect of burning and soil moisture on Organic C content (in kg/ha) and EC 1:5 (microS/cm)









Fig. 3 Biplot diagram of the first and second axes obtained from the principal components analysis (PCA) of the soil samples (above) and soil properties (down) data.